

1 **Temporal interactions among throughfall, type of canopy and thinning drive**
2 **radial growth in an Iberian mixed pine-beech forest**

3

4 Adrián Cardil¹, J. Bosco Imbert^{1*}, J. Julio Camarero², Irantzu Primicia¹ and Federico
5 Castillo¹

6

7 ¹ Depto. Ciencias del Medio Natural, Universidad Pública de Navarra, Campus de
8 Arrosadía, Pamplona, Navarra, 31006, Spain.

9 ² Instituto Pirenaico de Ecología (IPE-CSIC), Avda. Montañana, 1005, 50192 Zaragoza,
10 Spain

11

12 * Corresponding author:

13 Dr. J. Bosco Imbert

14 Dep. Ciencias del Medio Natural, Universidad Pública de Navarra,

15 Campus de Arrosadía, Pamplona, Navarra, 31006, Spain.

16 Tel: +34 948 16 98 59

17 Fax: +34 948 16 89 30

18 E-mail: bosco.imbert@unavarra.es

19

20 **Abstract**

21 Many factors can influence tree growth over time such as different forest management
22 practices, climate or tree-to-tree interactions, especially in mixed forests. We show in
23 this work how the temporal growth patterns for Scots pine and European beech depend
24 on thinning intensity (0 %, 20 % and 40 % extraction of basal area), canopy type (pine–
25 beech vs. pine patches), throughfall and their interactions. To fulfill this objective we
26 monitored radial growth of both species using band dendrometers during a 6-year long
27 period including two very dry years. Temporal growth patterns differed between both
28 species. Whereas Scots pine showed two main peaks of growth in May-June and
29 October, European beech mainly grew from May to early September when throughfall
30 was very limited. Effects of thinning on growth generally increased for both species
31 during dry periods both at the seasonal and annual scales. The treatment with 20 % of
32 thinning intensity was the most effective at the annual scale for enhancing growth of
33 both species. However, increases in growth due to thinning were much higher in beech
34 than in pine and lasted longer. Thinning effects on pine were higher in mixed canopy
35 than in pure canopy and appeared to be modulated by throughfall. Global differences in
36 pine growth between canopy types as a function of throughfall increased during the
37 main growing season as beech canopy developed. Growth of Scots pine, but not that of
38 European beech, generally increased with throughfall which suggests that pine might be
39 more dependent for its growth on water from the soil surface layer while beech would
40 depend more on water from deeper soil layers. Our findings have implications to select
41 the most convenient thinning treatments and canopy type under a potential climate
42 change scenario characterized by warmer conditions, more severe droughts and less
43 throughfall.

44 **Keywords:** Dendrometer; Forest management; Growth; Mixed forest; Water
45 availability; Throughfall

46

47 1. Introduction

48 Forest researchers and managers consider that a conversion of pure coniferous forests
49 into a mixed conifer-hardwood forest, is greatly meaningful from an ecological and
50 economic point of view (Knoke et al., 2005; Spiecker, 2003). Mixed conifer-hardwood
51 woodlands may provide multiple benefits such as higher biodiversity, protection from
52 disease or stability (resilience) to disturbances including climate extremes as droughts
53 (Knoke et al., 2005; Pretzsch et al., 2015). Additionally, mixed forests may increase
54 stand productivity when inter-specific processes are not dominated by competition
55 (Kelty, 1992). Nevertheless, empirical data on these topics is often nonexistent in mixed
56 conifer-hardwood forests (Forrester and Tang, 2016), particularly long-term data are
57 needed to elucidate mechanisms underlying observed patterns.

58 The growth dynamics in mixed conifer-hardwood forests are often difficult to predict
59 due to a high number of interactions among coexisting tree species as well as to the fact
60 that resource availability and climatic conditions change temporally and from site to site
61 (Forrester and Tang, 2016). Although many studies have been carried out on the growth
62 of Scots pine (*Pinus sylvestris* L.) or European beech (*Fagus sylvatica* L.) forests
63 (Camarero et al., 1998; Ježík et al., 2011), relatively few scientific literature is available
64 about inter-specific processes in mixed Scots pine and European beech forests (but see
65 Pretzsch et al., 2015; Pretzsch et al., 2016 and references therein). This is one of the
66 most widespread and economically relevant mixed forest type in Europe (Pretzsch et al.,
67 2015b), with a high ecological amplitude spanning from mesic conditions in central
68 Europe up to drought-prone conditions in sub-Mediterranean sites (del Río et al., 2014;
69 Primicia et al., 2016, 2013). A more severe drought stress due to warmer conditions,
70 higher evapotranspiration rates and more frequent heat waves (Cardil et al., 2014) are
71 one of the major climatic concerns negatively impacting forest productivity and vitality,
72 particularly in the Mediterranean Basin (Camarero et al., 2015). Indeed, temperature
73 and water availability during the summer are the main climatic drivers of Scots pine
74 (Martínez-Vilalta et al., 2008; Sánchez-Salguero et al., 2015) and European beech
75 (Rozas et al., 2015) long-term growth in Mediterranean populations. In this respect,
76 ongoing long-term increases in drought frequency and intensity in the Iberian Peninsula
77 (Andreu et al., 2007; Candel-Pérez et al., 2012) are reducing the productivity and
78 increasing the incidence of canopy dieback and mortality in Scots pine (Camarero et al.,
79 2015; Gea-Izquierdo et al., 2014; Martínez-Vilalta and Piñol, 2002). In NE Spain

80 increased summer drought stress is impacting negatively on Scots pine growth, although
81 so far the magnitude of such punctual growth declines is not enough to counteract the
82 overall increase of growth during the 20th century (Martínez-Vilalta et al., 2008). In this
83 region, summer drought increase has even caused an upward shift (Peñuelas and Boada,
84 2003) and a dramatic growth decline of beech over the last half-century (Jump et al.,
85 2006). However, the beech populations are expanding across the mostly humid and
86 temperate areas of NW Spain, particularly in lowlands, and stand growth is expected to
87 show similar positive trends (Sánchez de Dios et al., 2016).

88 Drought-stress release of beech may be enhanced in mixed forests or stands relative to
89 pure ones (Mölder and Leuschner, 2014). However, mixing may not fully compensate
90 for lower field capacities or frequent lack of precipitation as shown by Metz et al.
91 (2016) when beech was admixed with Scots pine. Furthermore, mixing of Scots pine
92 and European beech may result in a reduction of throughfall (Primicia et al., 2013)
93 affecting tree growth negatively. Another strategy for minimizing forest drought
94 vulnerability in temperate forests (D'Amato et al., 2013; Elkin et al., 2015; Sohn et al.,
95 2016a) and also in drought-prone Mediterranean forests (Martín-Benito et al., 2010) is
96 thinning. Thinning has been recommended as a short-term solution to reduce drought
97 vulnerability by increasing resilience in mixed forests, albeit the initial enhancement of
98 resilience may be reversed as stands mature and bigger trees increase water demand
99 (D'Amato et al., 2013). Heavy thinning has been shown to improve resistance and
100 particularly resilience of Scots pine to drought events in Germany, but these benefits
101 diminish with time since the last intervention (Sohn et al., 2016a). In this case,
102 progressive decreases in growth recovery apparently resulted from decreased
103 throughfall and increased transpiration rather than from tree ageing. With regard to
104 beech, thinning can increase its growth during wet and also in dry years (Diaconu et al.,
105 2015; van der Maaten, 2013) but also affect growth negatively during dry periods by
106 reducing soil water availability layer (Geßler et al., 2004). Sohn et al., (2016a)
107 suggested that long-lasting effects on growth recovery are related to relatively larger
108 leaf area and fine-root biomass in thinned mixed stands. Consequently, they predicted
109 that European beech would recover faster from drought than Scots pine given its faster
110 expansion of crown and root systems. However, to our knowledge no single study has
111 dealt with the relationship between thinning intensity and growth response to drought in
112 mixed pine-beech forests. This is important because synergies between pine and beech

113 may result in different growth responses to thinning and drought relative to
114 monospecific stands. Furthermore, growth responses to thinning are typically studied at
115 an annual scale, but studying seasonal scales (Aldea et al., 2017; Primicia et al., 2013)
116 may be important to elucidate the mechanisms implied and the long-term effects on
117 growth. For instance, temporal patterns of rainfall and leaf phenology may interact and
118 modulate the effects of thinning on tree growth.

119 In this work we analyze whether temporal interactions among throughfall, canopy type
120 and thinning affected radial growth of Scots pine and European beech during a 6-year
121 long period (2009-2014) which included two very dry years. The research site, Aspurz,
122 is a thinned pine-beech forest located in the western Spanish Pyrenees, close to the
123 southern European limit of Scots pine and European beech. Previous research has
124 shown that throughfall is the most important growth limiting factor for Scots pines at
125 this site (Primicia et al., 2013). Aspurz can be considered as a transitional site between
126 the Eurosiberian and Mediterranean regions, respectively characterized by cool-wet and
127 warm-dry climate conditions, in which since 1920 mean annual temperatures have risen
128 $+0.020\text{ }^{\circ}\text{C year}^{-1}$ and water surplus has significantly decreased (González de Andrés et
129 al., 2017).

130 Linear mixed models are a valuable statistical tool for describing and predicting the
131 radial-growth patterns in relation to different factors such as thinning and canopy type
132 in mixed forests (Pretzsch et al., 2015). Therefore, these models were used in this study
133 in order to fulfill the following objectives: (1) To compare intra-annual and inter-annual
134 basal-area growth patterns of Scots pine and European beech in a mixed forest; (2) To
135 evaluate the effects of canopy type and thinning on Scots pine basal-area growth
136 dynamics, and of thinning on basal area growth dynamics of European beech; and (3)
137 To assess the effects of throughfall and its interactions with other factors (i.e., month,
138 year, type of canopy and thinning) or combinations of these factors on basal-area
139 growth dynamics of Scots pine and European beech.

140

141 2. Methods

142 2.1. Study area

143 The study area is located in Aspurz ($42^{\circ}42'31''\text{N}$, $1^{\circ}08'40''\text{W}$), a forest situated in the
144 western Spanish Pyrenees, Navarre. Plots are North-oriented, located at a mean altitude

145 of 642 m, a mean slope of 7 % and on a soil classified as Haplic Alisol (Blanco, 2004).
146 The climate is considered as a cold wet Mediterranean type with water deficit usually in
147 July and August with frequent frosts from winter to early spring. In the 1984-2014
148 period, mean annual precipitation was 921 mm and mean annual temperature 12.0 °C
149 (data from Navascués weather station located at 2.7 km from the plots, 42° 43' 06'' N,
150 1° 06' 55'' W, 615 m). The forest is an even-aged mixed stand dominated by *Pinus*
151 *sylvestris* L. (Scots pine), naturally regenerated in the mid-1960s after strip-like clear-
152 cutting, with a mean dominant tree height of 20.4 m in 2014, being one of the most
153 productive Scots pine forests in Spain (Primicia et al., 2016). *Fagus sylvatica* L.
154 (European beech) is the second most abundant tree species, with a mean crown cover of
155 ca. 38 % and mean age of 35–40 years. Most beech trees were codominant or dominant,
156 although they covered all strata from suppressed to dominant ones.

157 2.2. Experimental design

158 A complete randomized block design (Andrew, 1986) was used in this study, including
159 nine plots (30 m x 40 m) installed by the Servicio Forestal del Gobierno de Navarra.
160 The experimental design can be thus specified as a split-plot (von Ende, 2001). Three
161 different thinning intensities (3 thinning intensities with 3 replicates) were applied in
162 November 1999 (0 %, 20 % and 30 % basal area removed) and March 2009 (0 %, 20 %
163 and 40 %) in the plots and in a buffer zone of 5-10 m around them. The first thinning
164 was carried out removing mainly suppressed or intermediate trees, and some dominant
165 or codominant trees with malformed stems. However, during the second thinning,
166 mainly subdominant or dominant trees were removed. The highest thinning intensity
167 was increased up to the 40 % of the basal area during the second thinning in accordance
168 to the silvicultural trends applied on similar surrounding stands at that moment.
169 Similarly, following the silvicultural guidelines applied in Navarre on mixed stands
170 with beech as secondary species, only Scots pine trees were thinned.

171 Within each plot, two discontinuous subplots can be distinguished: mixed beech-
172 pine and pure pine subplots (hereafter abbreviated as MC and PC sub-plots). To
173 establish the subplots, first we divided each plot into 300 2 m x 2 m quadrats by using
174 banderoles. Then, we delimited the projection on the ground of beech canopy taller
175 than 2 m by using plastic bands. Lastly we used plot maps with georeferenced trees and
176 a grid of 2 m x 2 m quadrats to draw the contours of each type of subplot. In each plot,
177 we randomly selected three Scots pines in PC sub-plots, three Scots Pines and three

178 beeches in MC subplots. Only dominant or codominant trees were selected. Six
179 treatments are thus determined for pines (P) and three for beeches (B): BMT0 (beeches
180 in unmanaged mixed beech-pine), PMT0 (pines in unmanaged mixed beech-pine),
181 PPT0 (pines in unmanaged pure pine), BMT20 (beeches in mixed beech-pine in
182 moderate thinning), PMT20 (pines in mixed beech-pine in moderate thinning), PPT20
183 (pines in pure pine in moderate thinning), BMT40 (beeches in mixed beech-pine in
184 severe thinning), PMT40 (pines in mixed beech-pine in severe thinning) and PPT40
185 (pines in pure pine in severe thinning).

186 2.2.1. Seasonal dynamics of radial increment

187 In total, 54 Scots pine trees (6 per plot) and 27 beeches (3 per plot) were selected to
188 assess seasonal patterns of radial increments considering thinning intensities, canopy
189 type, throughfall, date and their interactions. Band manual dendrometers (DB20
190 Environmental Measuring Systems, Brno, Czech Republic) were placed at about 1.3 m
191 height around the stem after removing carefully the rhytidome (dead bark) to quantify
192 cumulative radial growth. We read them from March 2009 to November 2014 with a
193 biweekly frequency from March to June, when most growth occurs in the study species
194 (Camarero et al., 2010) and monthly from July to November. To avoid biases during
195 recording, all measures were always taken before midday, and they were never taken
196 after a rainy day. Finally, the cumulative radial growth data was converted to basal area
197 increment assuming a circular shape of the stems. These data were subtracted from the
198 previous data measured and the result divided by the days elapsed between
199 measurements to calculate growth rates as daily basal-area increment (hereafter BA_{Ir}, in
200 cm² day⁻¹).

201 2.3. Throughfall

202 We measured throughfall monthly using 54 collectors placed in the plots (3 per subplot)
203 in mixed and pure areas. Each collector consisted of a plastic funnel (diameter 23.5 cm)
204 mounted onto a 25 L opaque polyethylene collection container. A filter of fiber-glass
205 coated with polyethylene (1.5 mm mesh size) was placed into the neck of the funnel to
206 avoid the entrance of coarse debris. Throughfall data were regressed with monthly
207 precipitation data obtained from the nearby Navascués station in the nine plots
208 considering mixed and pure patches within each plot (18 regressions) to estimate daily
209 throughfall in each container ($r^2 = 0.88 - 0.99$, $p < 0.001$).

210 2.4. Xylogenesis

211 The cambium generates tracheids passing through different developmental stages. The
212 xylem differentiation process is called xylogenesis and can be studied by repeatedly
213 taking and observing wood samples (Fukuda, 1996). Two dominant or codominant
214 Scots pines ($n = 18$) and one European beech per plot ($n = 9$) were randomly selected
215 for this purpose. Xylogenesis was monthly monitored from April to December 2011 by
216 sampling wood microcores (2 mm in diameter, 1 - 2 cm in length) at a height of 1.3 m
217 following a spiral around the stem using a Trephor borer (Rossi et al., 2006).
218 Microcores were always collected in areas apart from dendrometers to avoid distorting
219 dendrometer records. The methods of processing, sectioning and staining micro-core
220 wood sections are described with more detail elsewhere (Antonova et al., 1983;
221 Primicia et al., 2013). For both species, we measured the intra-annual radial growth in
222 wood microcores. Microcore data were used to validate the dendrometer recordings and
223 to justify its use for representing radial growth and not swelling-shrinking dynamics of
224 the trunk (Sheil, 2003). Lastly, correlations between radial-increment measurements
225 taken by dendrometers and those obtained from microcores were calculated to justify
226 the accuracy and reliability of the growth measurements recorded by dendrometers.

227

228 2.5. Statistical analyses

229 All models and the statistical analyses were carried out using R 3.2.4 (R core
230 development team, 2017). Linear mixed-effects models were used to assess the effects
231 of type of canopy (pure pine, mixed beech-pine), thinning intensity (0%, 20%, 40%),
232 throughfall (continuous variable), month (April-October), year (2009-2014) and their
233 interactions on Scots pine and beech growth (BAIr) using the “nlme” package (Pinheiro
234 and Bates, 2000), following a split-plot design with repeated measures. Thinning
235 intensity, canopy type, throughfall, month and year were included in the model as fixed
236 effects, and tree nested in plot, as random effect. Additionally, we included an
237 autoregressive correlation structure of first order to account for the repeated measures
238 on the same tree, and a variance structure, if the residual spread differed per month
239 (Pinheiro and Bates, 2000). Thus, the proposed model was stated as follows:

240

$$241 \quad \text{BAIr} = I + \text{thinning} * \text{canopy} * \text{throughfall} + \text{month} + \text{year} + \text{tree:plot} + \\ 242 \quad \text{corAR}(\text{tree:plot}) + \text{varIdent}(\text{month}) \quad (1)$$

243

244 where: BAIr represents the daily basal area increment measured with dendrometers; I is
245 the intercept; thinning, canopy and throughfall represent the study treatments; month
246 and year are the sampled month and year for which BAIr has been estimated; tree:plot, a
247 random term specifying the effect of individual trees nested within each plot;
248 corAR(tree:plot), the autoregressive correlation structure to account for the repeated
249 measures on each tree within each plot; varIdent (month), variance structure to account
250 for the different spread of residuals per month. **Throughfall was included in the model
251 as a continuous variable considering the sum of 10 days before each dendrometer
252 sampling date to take into account potentially lagged climate-growth responses
253 (Camarero et al., 1998; Primicia et al., 2013).**

254 We compared nested models with and without the random and the correlation and
255 variance structures to analyse their appropriateness calculating the likelihood ratio test
256 using the restricted maximum likelihood estimation procedure. Similarly, the
257 significance of the fixed effects was determined by comparing nested models with and
258 without the fixed terms, but using the maximum likelihood estimation procedure (Zuur
259 et al., 2009). When interpretable, significant three- and two-way interactions were
260 described. If there was no evidence for dependence between factors, main factors were
261 analyzed (Underwood, 1997).

262 **Throughfall intervals were delimited in the graphs in order to highlight the growth
263 patterns, to show significant differences in growth between different levels of thinning,
264 canopy type or month for each throughfall interval by using Helmert contrasts
265 (Chambers and Hastie, 1992), and to describe the results more clearly. Five throughfall
266 intervals ([0-10], [10-20], [20-40], [40-80], >80 mm) were selected by using histograms
267 to provide the most uniform distribution of measurements.**

268

269 3. Results

270 3.1. Relationships between radial-increment dynamics and xylogensis

271 Similar values of radial increment for Scots pine and European beech were recorded
272 using dendrometers and xylogensis during 2011 as shown in Figure 1. Measurements
273 between the two methods were highly correlated ($r=0.905$ for Scots pines and $r=0.951$
274 for European beeches). Therefore, measurements of radial growth based on
275 dendrometers allow interpreting and discussing adequately the results of this work.

276 3.2. Annual and seasonal trends of tree growth

277 Month and year significantly influenced BAIR in both Scots pine and European beech
278 (Table 1, $p < 0.001$). Intra-annual growth patterns differed between both species (Figure
279 2). BAIR in Scots pine showed two main peaks in the year, a major peak in late spring
280 from the beginning of May (Day of the Year, DOY 105) to the middle of July (DOY
281 192), centered in mid May and June, and the other in early autumn (October,
282 specifically from DOY 285 to 303). However, the temporal growth pattern was different
283 for European beeches with a peak from late spring (DOY 130) to early autumn (DOY
284 255) from May to early September, centered in June and July (Figure 2), being
285 throughfall very limited in the latter. The highest growth rates were observed in late
286 May and late June in the case of Scots pine ($2.4 \text{ cm}^2 \text{ day}^{-1}$) and European beech (7.0
287 $\text{cm}^2 \text{ day}^{-1}$), respectively. The growing period during the six study years took around 105
288 days considering the two main peaks for Scots pines and 125 days for European beech.
289 In 2011, the onset of xylogenesis was 60-90 and 122-152 DOY, and the cessation
290 occurred in 270 and 276 DOY for Scots pine and European beech, respectively.

291 3.3. Type of canopy and thinning

292 A significant interaction ($p = 0.011$) between type of canopy and month was found for
293 Scots pine (Table 1). Thus, except for August, BAIR in Scots pines was significantly
294 higher in pure patches during the months when most growth occurred (May to July and
295 October) but no significant differences between pure and mixed patches were found at
296 low BAIR values (i.e., April and September) (Figure 3a). On an annual basis, BAIR in
297 pure pine patches was on the average 22.1 % (range 17.3 to 28.2%) higher than in
298 mixed pine-beech patches, being these differences significant ($P < 0.05$) for all years
299 excepting 2012 (Figure 3b).

300 Thinning interacted significantly with month and year in Scots pine and with month in
301 European beech (Table 1, Figure 4). Thinning effects on Scots pine were only
302 significant ($P < 0.01$) in May (Figure 4a) but for European beech significant differences
303 ($P < 0.01$) were observed from May to September (Figure 4c). In Scots pine, growth
304 differences in moderately (20% of removed basal area) and severely (40% of removed
305 basal area) thinned plots relative to unthinned plots diminished from almost 50% in
306 April to about 10% in June. Then, in July and August differences remained at 10% for
307 moderate thinning but increased to 25-30 % in severe thinning (Figure 4a). In contrast,

308 beech differences relative to unthinned plots were almost constant from May to
309 September (around 40%) in severe thinning, being the effects of moderate thinning
310 much higher especially from June (80%) to August (120%) (Figure 4c). On an annual
311 basis, significant differences for pine occurred only in 2011 ($P < 0.001$) and for beech
312 from 2009 to 2012 ($P < 0.001$), and in 2013 ($p < 0.01$) and 2014 ($P < 0.05$) (Figures
313 4b,d). Growth of pine in thinned plots relative to that of unthinned plots showed a
314 unique positive peak in 2011 for both moderate (58%) and severe thinning (40%); by
315 2014, five years after thinning, the effect diminished until 20% and -5% under severe
316 and moderate thinning, respectively (Figure 4b). In contrast, beech exhibited a
317 progressive growth increase under moderate thinning from 2009 (83%) until 2012
318 (123%), to decrease afterwards to a minimum in 2014 (60%); under severe thinning
319 relative growth increased from 20% in 2009 to a plateau from 2010 to 2012 (around
320 48%) decreasing subsequently to 24% in 2014 (Figure 4d).

321 3.4. Throughfall influences tree growth dynamics

322 Throughfall significantly influenced BAIR in both Scots pine and European beech and it
323 was included in most of the significant interactions with other variables (Table 1).

324 3.4.1. Scots pine

325 The interaction between throughfall and type of canopy changed monthly as no
326 significant effects of canopy on BAIR at different throughfall levels were found in April
327 and November, but significant effects were found for the other months, particularly
328 during the months with highest tree growth ($P = 0.017$; Table 1; Figure 5). Thus, global
329 canopy effects (pure > mixed) on pine growth increased from May to July. Significant
330 effects between canopy types were found in May, June (0-10 mm and 10-20 mm) and
331 July (0-10 mm and 20-40 mm). It is also interesting to point out that no canopy effects
332 on BAIR were detected above 20-40 mm.

333 The interaction between throughfall and thinning changed depending on the type of
334 canopy (Figure 6). Thus, in pure patches, thinning significantly affected BAIR at 0-10
335 mm in pure (0% > 20% = 40%, $P = 0.01$) and mixed patches (20% > 0%, 20% = 40%,
336 0% = 40%, $P = 0.05$), but the effects were relatively small. However, thinning did
337 significantly ($P = 0.001$) affect BAIR in mixed patches at throughfall intervals 20-40

338 mm (20% > (40% = 0%)) and 40-80 mm ((20% =40%) > 0%) being the effects bigger
339 than at 0-10 mm .

340 The interaction between thinning and throughfall also changed monthly ($P < 0.046$;
341 Table 1) and annually ($P < 0.001$; Table 1). The monthly change is not shown as the
342 pattern is difficult to interpret. Significant relationships between throughfall and
343 thinning were detected in 2010 and 2013 (40-80 mm throughfall) and especially in
344 2011, the driest year of the study period (10-20 mm, 20-40mm and 80-160 mm)
345 (Appendix, Figure A.1a). Normally, higher values of throughfall were linked to greater
346 growth for Scots pine, in most years except for 2011 (Table 1, throughfall * year, $P <$
347 0.001 , Figure A.1a), and especially during the growing season (i.e., May-July) (Table 1,
348 throughfall * month, $P < 0.01$; Figure 7). Particularly, BAIR increased concomitantly
349 with throughfall in May, June, August and October, whereas generally no matching was
350 found in relation to throughfall in April, July,September and November.

351 3.4.2. European beech

352 Unlike for pine, the interaction between throughfall, thinning and year was not
353 significant for beech ($P = 0.987$; Table 1). However, significant thinning effects were
354 observed in 2011 (0-10, 10-20 mm) and 2012 (0-10, 10-20, 20-40 mm), the two driest
355 years, and in 2010 (40-80 mm) and 2013 (20-40 mm) (Figure A.1b). Also differently to
356 pine, monthly BAIR generally did not increase concomitantly with throughfall , being
357 the interaction between throughfall and month not significant ($P = 0.309$; Table 1;
358 Figure 7B).

359 3.4.3. Canopy and thinning influence throughfall

360 Moderate and severe thinning presented the highest values of throughfall at annual scale
361 (Appendix, Figure A2.) . Significant differences were found in the mean annual
362 amounts of throughfall as a function of canopy type ($P < 0.01$). Containers placed in
363 mixed patches received less amount of throughfall than those in pure patches, especially
364 in unmanaged mixed beech-pine plots. Thinning increased average throughfall only in
365 mixed patches, 29.2% and 21.5% in moderate and severe thinning, respectively.

366 4. Discussion

367 4.1. Temporal and seasonal trends of tree growth

368 Tree growth and cambial resumption in both Scots pine and European beech started
369 from April to May coinciding with a raise in air temperatures and with enough soil
370 humidity as occur in other boreal, temperate and Mediterranean forests (Camarero et al.,
371 2010; Čufar et al., 2008; Deslauriers et al., 2008; Schmitt et al., 2004). However,
372 differences in daily basal area increments patterns were found between both species
373 which presented the main growth peaks in May (pine) and June (beech) (Fig. 2). These
374 different growth peaks agree with xylogenesis studies carried out in nearby sites
375 (Martinez del Castillo et al., 2016). Scots pine exhibited a sharp reduction in growth in
376 mid-July, similarly to a previous study at this site (Primicia et al., 2013), whereas
377 European beech showed high growth until early September. This growth of beech
378 during summer might be related in part to the creation of parenchyma as the temperature
379 was raised (Morris et al., 2016). The reduction in BAIr in Scots pine matched with the
380 summer drought (i.e., low amounts of throughfall and high evapotranspiration rates) and
381 was probably caused by a decline in the cambial activity (Camarero et al., 2010, Gruber
382 et al., 2010), and possibly stem shrinkage (Zweifel et al., 2001). Relatively high growth
383 of beech during the summer dry period, as compared with pine, appears to be caused by
384 tapping of water by beech roots from deep soil horizons despite fine roots of pine and
385 beech reach highest densities at the upper soil layers (Curt and Prévosto, 2003).
386 However, as stands mature beech roots become predominant throughout the soil profile.
387 This seems to be the case at our experimental plots as sampled beech trees are dominant
388 or codominant. Furthermore, beech fine roots may shift deeper within the soil profile in
389 response to drought (Meier and Leuschner, 2008), and even show highest surface area
390 in deeper soil horizons in mixed stands (Bolte and Villanueva, 2006). Taken altogether
391 these patterns would explain the high growth rates of beech during the dry period linked
392 to its greater capacity to uptake soil water, but more research needs to be carried out to
393 link root functioning with wood formation. In contrast, Martinez et al. (2016) and
394 Michelot et al. (2012), found a severe reduction of beech cambial activity in mid-
395 August. A second growing peak was found for Scots pine from September to October;
396 and this bimodal pattern has been previously observed by analysing wood microcores in
397 the study site (Primicia et al., 2013) and elsewhere (Linares et al., 2009). However,
398 under colder conditions, Scots pine, that has a Euro-Siberian distribution, may show a
399 unimodal pattern (Camarero et al., 1998) suggesting an interaction of climate with other
400 factors to drive xylogenesis plasticity. In this context, unimodal patterns have been
401 reported for European beech, which has mainly an Euro-Siberian distribution.

402

403 4.2. Effects of canopy type and thinning on tree growth dynamics

404 Beech presence in Aspurz has caused an increase in average yearly reduction of BAIR
405 for Scots pine in mixed relative to pure patches from 17% in 2007-2008 (Primicia et al.,
406 2013) to 27% in 2014. This reduction is expected to increase as more beeches reach the
407 upper part of the canopy. Scots pine growth reduction resulting from competition with
408 beech has also been documented in other Pyrenean mixed forests (del Río et al., 2014).
409 The fact that this growth reduction pattern occurred from May to October suggests that
410 Scots pine and beech may be competing for different types of resources (i.e., light,
411 water and nutrients) or a combination of resources along the year. Competition may be
412 particularly intense in May when Scots pine shows its highest growth and sprouting and
413 accelerated expansion of beech leaves occurs, and in June when pines reach its second
414 highest growth value and beech its maximum growth (Fig. 2). During this period
415 competition for water and nutrients needed for shoot, leaf and flower production may be
416 important. Indeed, significantly lower soil N-NH₄⁺ content in mixed than in pure
417 patches was detected in the same plots in 2007 and 2008 which suggests higher
418 ammonium uptake in mixed patches due to beech presence (Primicia, 2012). Beech may
419 have also outcompete pine by decreasing throughfall in mixed patches by 11.4% during
420 the leafless season and 20.5 % during the leafed season (Primicia, 2012). This effect on
421 growth was apparently higher in July and August, the driest months, when throughfall
422 reduction was higher (see figure 2 and A.1.). On a longer timescale, light was probably
423 the most limiting factor for suppressed pine trees growing nearby dominant or
424 codominant beech trees, whose leaves sprouted in April-May and were fully expanded
425 in September (Arretxe, 2010).

426

427 Thinning usually leads to improved growth in the short term (Elkin et al., 2015;
428 Linares et al., 2009; Primicia et al., 2016), an effect which can be explained by an
429 increase in soil water and nutrients availability to competing trees (Blanco et al., 2005),
430 and by an improved growth and photosynthetic capacity due to the increment of the
431 foliar mass of the crown (Aussenac, 2000). Absolute and relative increases in growth
432 due to thinning were much higher in beech than in Scots pine and lasted longer (Fig. 4).
433 This is not surprising given the fact that cover of beech at this site is increasingly
434 expanding relative to that of pine (Primicia et al., 2013) by taking advantage of its high

435 efficiency in space occupation (Pretzsch and Schütze, 2005). Additionally, based on
436 mean temperature and annual precipitation for Aspurz, conditions appear to be more
437 favorable for European beech than for Scots pine (Pretzsch et al., 2016).

438

439 Scots pine and beech showed different monthly and annual responses to thinning
440 (Fig. 4, Table 1). Positive growth differences between thinned and unthinned plots for
441 both species increased in July and August, the two driest months, indicating the efficacy
442 of thinning to mitigate drought stress. However, while severe thinning allowed Scots
443 pine to maintain relatively higher growth rates during the drought period, in the case of
444 beech the highest growth rates were achieved with moderate thinning. Wind speed
445 within the stand and evaporation may have increased following tree reduction, more so
446 at severe thinning, enhancing transpiration rates (Aussenac, 2000). Additionally, larger
447 crown surface area (Sohn et al., 2016b) and increases in understory vegetation cover
448 (unpublished results) under severe thinning may have also contributed to increase
449 transpirational demand. This would explain the more positive effect of moderate
450 thinning on beech growth, as less transpiration in this treatment than under severe
451 thinning could have resulted in a water balance more favorable. In this context, thinning
452 intensity probably did not affect so much between different thinning intensities in
453 transpiration of Scots pine, a drought-avoider species (Zweifel et al., 2001). Pine
454 probably reduced its stomata conductance at critical periods during the daytime as
455 indicated by low relative growth rates because of low surface soil water potential
456 (unpublished results). Consequently, severe thinning might have been more effective
457 than moderate thinning to mitigate growth reduction especially if root system
458 development of pines increased (Kneeshaw et al., 2002) and/or average fine-root depth
459 decreased (Giuggiola et al., 2016) with thinning intensity. The latter would be a
460 response to increased soil water availability in shallower soil depths resulting from an
461 increase in throughfall with thinning intensity

462

463 On an annual basis growth in thinned plots relative to that of unthinned plots showed a
464 positive peak in 2011 under moderate and to a less extent under severe thinning for
465 Scots pine, and a positive maximum in 2011 and 2012 under moderate thinning for
466 beech. These were the driest years during the study period, especially 2011 (663 mm of
467 rainfall; 1,042 mm of mean in 2009-2014 period; data from Navascués weather station;
468 Appendix, Figure A.3.). Therefore, thinning also mitigated growth reduction during

469 drought of Scots pine and European beech at an annual scale as it has been shown for
470 Scots pine in Switzerland (Giuggiola et al., 2013) and Central Europe (Sohn et al.,
471 2016a) and for beech in southwestern Germany (van der Maaten, 2013). Thus, if mean
472 annual temperature continues increasing and water surplus continues decreasing over
473 the long term at this region (González de Andrés et al., 2017), a 20 % thinning intensity
474 treatment could be the more efficient management option in terms of maximizing
475 growth.

476

477 4.3. Throughfall influences growth dynamics

478 Monitoring both intra-annual and inter-annual growth dynamics is a necessary approach
479 for characterizing the tree growth responses to different forest management treatments
480 such as thinning intensity and canopy type (Michelot et al., 2012). However, to
481 elucidate some mechanisms underlying the observed patterns it can be very relevant
482 looking at throughfall changes in relation to these treatments.

483 Global differences in pine BAIR between canopy types (pure > mixed, see Fig. 5) as a
484 function of throughfall increased from May to July, when most pine growth occurred.
485 Shoot elongation and leaf expansion of beech during this period apparently resulted in
486 less light (Pretzsch et al., 2016) and water (Primicia, 2012; Primicia et al., 2013)
487 reaching pine needles and roots, respectively, which could have lead to reduced growth
488 in mixed patches. Significant differences between canopy types were mainly observed
489 at low throughfall (0-10 mm) due in part to lower growth variability, and to a less extent
490 at medium flow (20-40 mm). No significant differences between canopy types were
491 found at high throughfall (40-80 mm, 80-160 mm) suggesting a throughfall threshold to
492 trigger these differences. These patterns may imply an increase in growth differences
493 between canopy types if the frequency of high throughfall events decreases, and that of
494 low-medium throughfall events increases. Furthermore, unlike in pure-canopy patches,
495 rainfall interception in mixed-canopy patches caused the reduction of maximum
496 throughfall input (i.e., the disappearance of throughfall interval 80-160 mm) into the
497 soil in May and October, and consequently a direct reduction of pine growth occurred
498 given the positive relationships between growth and throughfall amounts (see Fig. 7).
499 This pattern may strengthen as more beeches reach the upper canopy level, contributing
500 further to growth differences between canopy types. However, growth reduction over
501 time in mixed patches in response to decreasing throughfall probably will increase faster

502 in pine than in beech. This is because, at this site, water use efficiency (iWUE, i.e. the
503 ratio of assimilated carbon to water lost through stomata) of beech has increased more
504 (ca +22%) than that of pine (ca +12%) from 1980 to 2013 apparently as an adaptation to
505 increasing dryness (González de Andrés et al., 2017).

506 Although the double interaction thinning* canopy and the triple interactions
507 thinning*canopy* month or year were not significant for pine growth, the interaction
508 thinning* canopy changed significantly as a function of throughfall. This effect was
509 especially noticeable in mixed patches with throughfall of 20-40 mm and 40-80 mm in
510 which average growth was higher under moderate thinning than in severely thinned or
511 unthinned plots. In other words, the throughfall changed the thinning effects in mixed
512 patches but not in pure patches. This effect may have been caused by nutrients,
513 especially nitrates (Primicia, 2012), washed away by throughfall from the beech canopy
514 and the forest floor which is thicker in mixed patches. The higher effect of thinning on
515 pine growth in mixed patches under moderate thinning might be related to the lower
516 basal area of beech at these plots (Primicia et al., 2016) which would have resulted in
517 less competition on pines. A similar pattern was observed for the interaction between
518 thinning and type of canopy for throughfall (Figure A.2) indicating that the above
519 response of pine growth to thinning in mixed patches can also be generated by
520 differences in throughfall amounts among thinning intensities. Furthermore, in normal
521 or wet years significant thinning effects on pine and beech as a function of throughfall
522 were scarce. However, their frequency considerably increased in 2011 and 2012 in
523 beech for low-medium throughfall, and in 2011 in pine for low, medium and high
524 throughfall. The 2011 and 2012 years were especially dry, therefore these patterns
525 confirm that thinning treatments were effective to reduce some of the negative effects of
526 drought on growth, and that the main limiting resource during these years was water.

527 Growth of Scots pine generally increased with throughfall, especially in those months
528 when BA_{Ir} was higher. This confirms the role played by precipitation to enhance radial
529 growth in similar forests subjected to Mediterranean seasonal droughts (Bogino et al.,
530 2009; Camarero et al., 2010; Gutiérrez, 1989; Primicia et al., 2013)(Bogino et al., 2009;
531 Camarero et al., 2010; Gutiérrez, 1989; Primicia et al., 2013). However, monthly beech
532 growth was not generally related to throughfall. These two distinct patterns suggest that
533 pine might be more depending for its growth on water recently incorporated into the soil
534 (i., soil surface layer) while beech might depend relatively more on water from previous

535 rainfall events (i.e., deeper soil layers). These difference strategies could be further
536 investigated analyzing the isotope discrimination of oxygen and hydrogen in xylem and
537 soil water to identify the isotopic signature of water sources.

538 5. Conclusions

539 Scots pine and beech showed bimodal and unimodal temporal growth patterns,
540 respectively, but the spring growth peak was the most important in both species The
541 relatively high growth of beech during the summer dry period, as compared with pine,
542 could be explained by tapping of water by beech roots from deep soil horizons.
543 Competition between Scots pine and beech appeared to be high in May and June
544 coinciding with their respective peaks of growth. As beech competition for light
545 intensifies, pine growth is progressively diminishing in mixed patches relative to pure
546 patches. Absolute and relative increase in growth due to thinning were much higher in
547 beech than in Scots pine, providing to beech a high efficiency in space occupation in
548 sites with favorable growing conditions (e.g., drought stress). Differences in growth
549 between thinned and control plots for both species increased during the driest months
550 and years indicating the efficacy of thinning to mitigate drought stress. On an annual
551 basis moderate thinning appeared to be the most efficient treatment to maximize
552 growth. Finally, interactions between throughfall and the analyzed factors explain
553 mechanisms underlying pine growth reduction in mixed patches relative to pure
554 patches, differences in thinning effects on pine growth in pure and mixed patches,
555 higher efficiency of thinning during drought periods, and why beech trees were capable
556 of growing during the dry months.

557 6. Acknowledgements

558 We are thankful to the Public University of Navarre for supporting this study through a
559 grant to fund Cardil's Postdoctoral studies. This study has been carried out within the
560 framework of the projects AGL2006-08288 and AGL2009-11287 financed by
561 Ministerio de Economía y Competividad (Spain). Camarero JJ acknowledges the
562 support of the "Fundiver" Spanish project (Spanish Ministry of Economy and
563 Competitiveness, project CGL2015-69186-C2-1-R).

564

- 566 Aldea, J., Bravo, F., Bravo-Oviedo, A., Ruiz-Peinado, R., Rodríguez, F., del Río, M.,
567 2017. Thinning enhances the species-specific radial increment response to drought
568 in Mediterranean pine-oak stands. *Agric. For. Meteorol.* 237, 371–383.
569 doi:<http://dx.doi.org/10.1016/j.agrformet.2017.02.009>
- 570 Andreu, L., Gutiérrez, E., Macias, M., Ribas, M., Bosch, O., Camarero, J.J., 2007.
571 Climate increases regional tree-growth variability in Iberian pine forests. *Glob.*
572 *Chang. Biol.* 13, 804–815. doi:10.1111/j.1365-2486.2007.01322.x
- 573 Andrew, I., 1986. Simple experimental design for forestry trials. FRI Bull 71. Forest
574 Research Institute, Rotorua, New Zealand.
- 575 Antonova, G., Shebeko, V., Malyutina, E., 1983. Seasonal dynamics of cambial activity
576 and tracheid differentiation in the stem of Scots pine. *Chem. wood* 1, 16–22.
- 577 Arretxe, A., 2010. Influencia de las claras y el tipo de dosel en la morfología y la
578 composición química foliar de pino silvestre (*Pinus sylvestris*) y haya (*Fagus*
579 *sylvatica*) en un bosque mixto del Pirineo navarro (marzo 2008-enero 2009).
580 Public university of Navarre.
- 581 Aussenac, G., 2000. Interactions between forest stands and microclimate:
582 Ecophysiological aspects and consequences for silviculture. *Ann. For. Sci.* 57,
583 287–301. doi:10.1051/forest:2000119
- 584 Blanco, J., 2004. La práctica de las claras forestales y su influencia en el ciclo interno de
585 nutrientes en dos bosques de pino silvestre de los Pirineos Navarros. Universidad
586 Pública de Navarra, Spain.
- 587 Blanco, J., Zavala, M., Imbert, J., Castillo, F., 2005. Sustainability of forest
588 management practices: evaluation through a simulation model of nutrient cycling.
589 *For. Ecol. Manage.* 213, 209–228.
- 590 Bogino, S., Fernández Nieto, M.J., Bravo, F., 2009. Climate effect on radial growth of
591 *Pinus sylvestris* at its southern and western distribution limits. *Silva Fenn.* 43, 609–
592 623.
- 593 Bolte, A., Villanueva, I., 2006. Interspecific competition impacts on the morphology
594 and distribution of fine roots in European beech (*Fagus sylvatica* L.) and Norway
595 spruce (*Picea abies* (L.) Karst.). *Eur. J. For. Res.* 125, 15–26. doi:10.1007/s10342-
596 005-0075-5
- 597 Camarero, J., Gazol, A., Sangüesa-Barreda, G., Oliva, J., Vicente-Serrano, S., 2015.
598 To die or not to die: early warnings of tree dieback in response to a severe drought.
599 *J. Ecol.* 103, 44–57.
- 600 Camarero, J.J., Guada, G., Sánchez-Salguero, R., Cervantes, E., 2016. Winter drought
601 impairs xylem phenology, anatomy and growth in Mediterranean Scots pine
602 forests. *Tree Physiol.* 1–14. doi:10.1093/treephys/tpw077
- 603 Camarero, J.J., Guerrero-Campo, J., Gutierrez, E., 1998. Tree-ring growth and structure
604 of *Pinus uncinata* and *Pinus sylvestris* in the Central Spanish Pyrenees. *Arct. Alp.*
605 *Res.* 30, 1–10.
- 606 Camarero, J.J., Olano, J.M., Parras, A., 2010. Plastic bimodal xylogenesis in conifers
607 from continental Mediterranean climates. *New Phytol.* 185, 471–480.
608 doi:10.1111/j.1469-8137.2009.03073.x
- 609 Candel-Pérez, D., Linares, J.C., Viñepla, B., Lucas-Borja, M.E., 2012. Assessing
610 climate–growth relationships under contrasting stands of co-occurring Iberian
611 pines along an altitudinal gradient. *For. Ecol. Manage.* 274, 48–57.
612 doi:<http://dx.doi.org/10.1016/j.foreco.2012.02.010>
- 613 Cardil, A., Molina, D.M., Kobziar, L.N., 2014. Extreme temperature days and their

614 potential impacts on southern Europe. *Nat. Hazards Earth Syst. Sci.* 14, 3005–
615 3014. doi:10.5194/nhess-14-3005-2014

616 Chambers, J.M., Hastie, T.J., 1992. Chambers, J. M. and Hastie, T. J., in: Hall/CRC, C.
617 and (Ed.), *Statistical Models in S.* p. 624.

618 Čufar, K., Prislan, P., De Luis, M., Gričar, J., 2008. Tree-ring variation, wood formation
619 and phenology of beech (*Fagus sylvatica*) from a representative site in Slovenia,
620 SE Central Europe. *Trees - Struct. Funct.* 22, 749–758. doi:10.1007/s00468-008-
621 0235-6

622 Curt, T., Prévosto, B., 2003. Rooting strategy of naturally regenerated beech in Silver
623 birch and Scots pine woodlands. *Plant Soil* 255, 265–279.
624 doi:10.1023/A:1026132021506

625 D'Amato, A.W., Bradford, J.B., Fraver, S., Palik, B.J., 2013. Effects of thinning on
626 drought vulnerability and climate response in north temperate forest ecosystems.
627 *Ecol. Appl.* 23, 1735–1742. doi:10.1890/13-0677.1

628 del Río, M., Condés, S., Pretzsch, H., 2014. Analyzing size-symmetric vs. size-
629 asymmetric and intra- vs. inter-specific competition in beech (*Fagus sylvatica* L.)
630 mixed stands. *For. Ecol. Manage.* 325, 90–98. doi:10.1016/j.foreco.2014.03.047

631 Deslauriers, A., Rossi, S., Anfodillo, T., Saracino, A., 2008. Cambial phenology, wood
632 formation and temperature thresholds in two contrasting years at high altitude in
633 southern Italy. *Tree Physiol.* 28, 863–871. doi:10.1093/treephys/28.6.863

634 Diaconu, D., Kahle, H.P., Spiecker, H., 2015. Tree- and stand-level thinning effects on
635 growth of European Beech (*Fagus sylvatica* L.) on a Northeast- and a Southwest-
636 facing slope in southwest Germany. *Forests* 6, 3256–3277. doi:10.3390/f6093256

637 Elkin, C., Giuggiola, A., Rigling, A., Bugmann, H., 2015. Short- and long-term efficacy
638 of forest thinning to mitigate drought impacts in mountain forests in the European
639 Alps. *Ecol. Appl.* 25, 1083–1098. doi:10.1890/14-0690.1

640 Forrester, D.I., Tang, X., 2016. Analysing the spatial and temporal dynamics of species
641 interactions in mixed-species forests and the effects of stand density using the 3-
642 PG model. *Ecol. Modell.* 319, 233–254. doi:10.1016/j.ecolmodel.2015.07.010

643 Fukuda, H., 1996. Xylogenesis: initiation, progression, and cell death. *Annu. Rev. Plant*
644 *Physiol. Plant Mol. Biol.* 47, 299–325.

645 Gea-Izquierdo, G., Viguera, B., Cabrera, M., Cañellas, I., 2014. Drought induced
646 decline could portend widespread pine mortality at the xeric ecotone in managed
647 mediterranean pine-oak woodlands. *For. Ecol. Manage.* 320, 70–82.
648 doi:http://dx.doi.org/10.1016/j.foreco.2014.02.025

649 Geßler, A., Keitel, C., Nahm, M., Rennenberg, H., 2004. Water Shortage Affects the
650 Water and Nitrogen Balance in Central European Beech Forests. *Plant Biol.* 6,
651 289–298. doi:10.1055/s-2004-820878

652 Giuggiola, A., Bugmann, H., Zingg, A., Dobbertin, M., Rigling, A., 2013. Reduction of
653 stand density increases drought resistance in xeric Scots pine forests. *For. Ecol.*
654 *Manage.* 310, 827–835. doi:http://dx.doi.org/10.1016/j.foreco.2013.09.030

655 Giuggiola, A., Ogée, J., Rigling, A., Gessler, A., Bugmann, H., Treydte, K., 2016.
656 Improvement of water and light availability after thinning at a xeric site: which
657 matters more? A dual isotope approach. *New Phytol.* 210, 108–121.
658 doi:10.1111/nph.13748

659 González de Andrés, E., Camarero, J.J., Blanco, J.A., Imbert, J.B., Lo, Y.-H., Sangüesa-
660 Barreda, G., Castillo, F.J., 2017. Tree-to-tree competition in mixed European
661 beech–Scots pine forests has different impacts on growth and water-use efficiency
662 depending on site conditions. *J. Ecol.* doi:10.1111/1365-2745.12813

663 Gruber, A., Stöbl, S., Veit, B., Oberhuber, W., 2010. Impact of drought on the temporal

664 dynamics of wood formation in *Pinus sylvestris*. *Tree Physiol.* 30, 490–501.

665 Gutiérrez, E., 1989. Dendroclimatological study of *Pinus sylvestris* L. in southern
666 Catalonia (Spain). *Tree-Ring Bull.* 49, 1–9.

667 Ježík, M., Blaženec, M., Štrelcová, K., Ditmarová, Ľ., 2011. The impact of the 2003-
668 2008 weather variability on intra-annual stem diameter changes of beech trees at a
669 submontane site in central Slovakia. *Dendrochronologia* 29, 227–235.
670 doi:10.1016/j.dendro.2011.01.009

671 Jump, A.S., Hunt, J.M., Peñuelas, J., 2006. Rapid climate change-related growth decline
672 at the southern range edge of *Fagus sylvatica*. *Glob. Chang. Biol.* 12, 2163–2174.
673 doi:10.1111/j.1365-2486.2006.01250.x

674 Kelty, M.J., 1992. Comparative productivity of monocultures and mixed-species stands,
675 in: Kelty, M.J., Larson, B.C., Oliver, C.D. (Eds.), *The Ecology and Silviculture of*
676 *Mixed-Species Forests: A Festschrift for David M. Smith*. Springer Netherlands,
677 Dordrecht, pp. 125–141. doi:10.1007/978-94-015-8052-6_8

678 Kneeshaw, D.D., Williams, H., Nikinmaa, E., Messier, C., 2002. Patterns of above- and
679 below-ground response of understory conifer release 6 years after partial cutting.
680 *Can. J. For. Res.* 32, 255–265. doi:10.1139/x01-190

681 Knoke, T., Stimm, B., Ammer, C., Moog, M., 2005. Mixed forests reconsidered: A
682 forest economics contribution on an ecological concept. *For. Ecol. Manage.* 213,
683 102–116. doi:10.1016/j.foreco.2005.03.043

684 Linares, J.C., Camarero, J.J., Carreira, J.A., 2009. Plastic responses of *Abies pinsapo*
685 xylogenesis to drought and competition. *Tree Physiol.* 29, 1525–1536.
686 doi:10.1093/treephys/tpp084

687 Martín-Benito, D., Del Río, M., Heinrich, I., Helle, G., Cañellas, I., 2010. Response of
688 climate-growth relationships and water use efficiency to thinning in a *Pinus nigra*
689 afforestation. *For. Ecol. Manage.* 259, 967–975.
690 doi:http://dx.doi.org/10.1016/j.foreco.2009.12.001

691 Martínez-Vilalta, J., López, B.C., Adell, N., Badiella, L., Ninyerola, M., 2008.
692 Twentieth century increase of Scots pine radial growth in NE Spain shows strong
693 climate interactions. *Glob. Chang. Biol.* 14, 2868–2881. doi:10.1111/j.1365-
694 2486.2008.01685.x

695 Martínez del Castillo, E., Longares, L.A., Gričar, J., Prislan, P., Gil-Pelegrín, E., Čufar,
696 K., de Luis, M., 2016. Living on the Edge: Contrasted Wood-Formation Dynamics
697 in *Fagus sylvatica* and *Pinus sylvestris* under Mediterranean Conditions. *Front.*
698 *Plant Sci.* 7, 370. doi:10.3389/fpls.2016.00370

699 Martínez-Vilalta, J., Piñol, J., 2002. Drought-induced mortality and hydraulic
700 architecture in pine populations of the NE Iberian Peninsula. *For. Ecol. Manage.*
701 161, 247–256. doi:http://dx.doi.org/10.1016/S0378-1127(01)00495-9

702 Meier, I.C., Leuschner, C., 2008. Belowground drought response of European beech:
703 fine root biomass and carbon partitioning in 14 mature stands across a precipitation
704 gradient. *Glob. Chang. Biol.* 14, 2081–2095. doi:10.1111/j.1365-
705 2486.2008.01634.x

706 Metz, J., Annighöfer, P., Schall, P., Zimmermann, J., Kahl, T., Schulze, E.-D., Ammer,
707 C., 2016. Site-adapted admixed tree species reduce drought susceptibility of
708 mature European beech. *Glob. Chang. Biol.* 22, 903–920. doi:10.1111/gcb.13113

709 Michelot, A., Simard, S., Rathgeber, C., Dufrière, E., Damesin, C., 2012. Comparing the
710 intra-annual wood formation of three European species (*Fagus sylvatica* , *Quercus*
711 *petraea* and *Pinus sylvestris*) as related to leaf phenology and non-structural
712 carbohydrate dynamics. *Tree Physiol.* 1033–1043. doi:10.1093/treephys/tps052

713 Mölder, I., Leuschner, C., 2014. European beech grows better and is less drought

714 sensitive in mixed than in pure stands: tree neighbourhood effects on radial
715 increment. *Trees* 28, 777–792. doi:10.1007/s00468-014-0991-4

716 Morris, H., Plavcová, L., Cvecko, P., Fichtler, E., Gillingham, M.A.F., Martínez-
717 Cabrera, H.I., McGlenn, D.J., Wheeler, E., Zheng, J., Ziemińska, K., Jansen, S.,
718 2016. A global analysis of parenchyma tissue fractions in secondary xylem of seed
719 plants. *New Phytol.* 209, 1553–1565. doi:10.1111/nph.13737

720 Peñuelas, J., Boada, M., 2003. A global change-induced biome shift in the Montseny
721 mountains (NE Spain). *Glob. Chang. Biol.* 9, 131–140. doi:10.1046/j.1365-
722 2486.2003.00566.x

723 Pinheiro, J., Bates, D., 2000. *Mixed-effects models in S and S-PLUS*. Springer, New
724 York.

725 Pretzsch, H., del Río, M., Ammer, C., Avdagic, A., Barbeito, I., Bielak, K., Brazaitis,
726 G., Coll, L., Dirnberger, G., Drössler, L., Fabrika, M., Forrester, D.I., Godvod, K.,
727 Heym, M., Hurt, V., Kurylyak, V., Löf, M., Lombardi, F., Matović, B., Mohren,
728 F., Motta, R., den Ouden, J., Pach, M., Ponette, Q., Schütze, G., Schweig, J.,
729 Skrzyszewski, J., Sramek, V., Sterba, H., Stojanović, D., Svoboda, M.,
730 Vanhellefont, M., Verheyen, K., Wellhausen, K., Zlatanov, T., Bravo-Oviedo, A.,
731 2015. Growth and yield of mixed versus pure stands of Scots pine (*Pinus sylvestris*
732 L.) and European beech (*Fagus sylvatica* L.) analysed along a productivity gradient
733 through Europe. *Eur. J. For. Res.* 134, 927–947. doi:10.1007/s10342-015-0900-4

734 Pretzsch, H., del Río, M., Schütze, G., Ammer, C., Annighöfer, P., Avdagic, A.,
735 Barbeito, I., Bielak, K., Brazaitis, G., Coll, L., Drössler, L., Fabrika, M., Forrester,
736 D.I., Kurylyak, V., Löf, M., Lombardi, F., Matović, B., Mohren, F., Motta, R., den
737 Ouden, J., Pach, M., Ponette, Q., Skrzyszewski, J., Sramek, V., Sterba, H.,
738 Svoboda, M., Verheyen, K., Zlatanov, T., Bravo-Oviedo, A., 2016. Mixing of
739 Scots pine (*Pinus sylvestris* L.) and European beech (*Fagus sylvatica* L.) enhances
740 structural heterogeneity, and the effect increases with water availability. *For. Ecol.*
741 *Manage.* 373, 149–166. doi:http://dx.doi.org/10.1016/j.foreco.2016.04.043

742 Pretzsch, H., Schütze, G., 2005. Crown Allometry and Growing Space Efficiency of
743 Norway Spruce (*Picea abies* [L.] Karst.) and European Beech (*Fagus sylvatica* L.)
744 in Pure and Mixed Stands. *Plant Biol.* 7, 628–639. doi:10.1055/s-2005-865965

745 Pretzsch, Forrester, D.I., Rötzer, T., 2015. Representation of species mixing in forest
746 growth models: A review and perspective. *Ecol. Modell.* 313, 276–292.
747 doi:10.1016/j.ecolmodel.2015.06.044

748 Primicia, I., 2012. Influence of thinning and canopy type on the internal nutrient cycling
749 and the secondary growth of *Pinus sylvestris* L. in a mixed forest in the Pyrenees.
750 Public University of Navarre, Spain.

751 Primicia, I., Camarero, J.J., Imbert, J.B., Castillo, F.J., 2013. Effects of thinning and
752 canopy type on growth dynamics of *Pinus sylvestris*: Inter-annual variations and
753 intra-annual interactions with microclimate. *Eur. J. For. Res.* 132, 121–135.
754 doi:10.1007/s10342-012-0662-1

755 Primicia, I., Imbert, J., Puertas, F., Traver, M. del C., Castillo, F.J., 2016. Influence of
756 thinning intensity and canopy type on Scots pine stand and growth dynamics in a
757 mixed managed forest. *For. Syst.* 25, e057.

758 R core development team, 2017. R 3.2.4.

759 Rossi, S., Anfodillo, T., Menardi, R., 2006. Trephor: A new tool for sampling
760 microcores from tree stems. *IAWA J.* 27, 89–97. doi:10.1163/22941932-90000139

761 Rozas, V., Camarero, J., Sangüesa-Barreda, G., Souto, M., García-González, I., 2015.
762 Summer drought and ENSO-related cloudiness distinctly drive *Fagus sylvatica*
763 growth near the species rear-edge in northern Spain. *Agric. For. Meteorol.* 201,

764 153–164.
765 Sánchez-Salguero, R., Camarero, J.J., Hevia, A., Madrigal-González, J., Linares, J.C.,
766 Ballesteros-Canovas, J.A., Sánchez-Miranda, A., Alfaro-Sánchez, R., Sangüesa-
767 Barreda, G., Galván, J.D., Gutiérrez, E., Génova, M., Rigling, A., 2015. What
768 drives growth of Scots pine in continental Mediterranean climates: Drought, low
769 temperatures or both? *Agric. For. Meteorol.* 206, 151–162.
770 doi:<http://dx.doi.org/10.1016/j.agrformet.2015.03.004>
771 Sánchez de Dios, R., Hernández, L., Montes, F., Sainz-Ollero, H., Cañellas, I., 2016.
772 Tracking the leading edge of *Fagus sylvatica* in North-Western Iberia: Holocene
773 migration inertia, forest succession and recent global change. *Perspect. Plant Ecol.*
774 *Evol. Syst.* 20, 11–21. doi:<http://dx.doi.org/10.1016/j.ppees.2016.03.001>
775 Schmitt, U., Jalkanen, R., Eckstein, D., 2004. Cambium dynamics of *Pinus sylvestris*
776 and *Betula* spp. in the northern boreal forest in Finland. *Silva Fenn.* 38, 167–178.
777 Sheil, D., 2003. Growth assessment in tropical trees: Large daily diameter fluctuations
778 and their concealment by dendrometer bands. *Can. J. For. Res.* 33, 2027–2035.
779 Sohn, J.A., Hartig, F., Kohler, M., Huss, J., Bauhus, J., 2016a. Heavy and frequent
780 thinning promotes drought adaptation in *Pinus sylvestris* forests. *Ecol. Appl.* 26,
781 2190–2205. doi:10.1002/eap.1373
782 Sohn, J.A., Saha, S., Bauhus, J., 2016b. Potential of forest thinning to mitigate drought
783 stress: A meta-analysis. *For. Ecol. Manage.* 380, 261–273.
784 doi:<http://dx.doi.org/10.1016/j.foreco.2016.07.046>
785 Spiecker, H., 2003. Silvicultural management in maintaining biodiversity and resistance
786 of forests in Europe-boreal zone: Case Finland. *J. Environ. Manage.* 67, 47–54.
787 doi:10.1016/S0301-4797(02)00187-1
788 Underwood, A.J., 1997. Experiments in ecology. Their logical design and interpretation
789 using analysis of variance. Cambridge University Press, Cambridge (UK).
790 van der Maaten, E., 2013. Thinning prolongs growth duration of European beech (*Fagus*
791 *sylvatica* L.) across a valley in southwestern Germany. *For. Ecol. Manage.* 306,
792 135–141. doi:<http://dx.doi.org/10.1016/j.foreco.2013.06.030>
793 von Ende, C., 2001. Repeated-measures analysis: growth and other time dependent
794 measures., in: *The Design and Analysis of Ecological Experiments*. Oxford
795 University Press, New York, pp. 134–157.
796 Zuur, A., Ieno, E., Walker, N., Saveliev, A., Smith, G., 2009. *Mixed effects models and*
797 *extensions in ecology with R*. Springer, New York.
798 Zweifel, R., Item, H., Häslter, R., 2001. Link between diurnal stem radius changes and
799 tree water relations. *Tree Physiol.* 21, 869–877. doi:10.1093/treephys/21.12-13.869
800
801

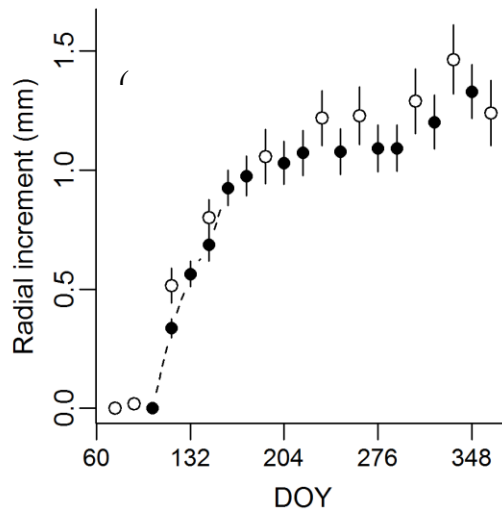
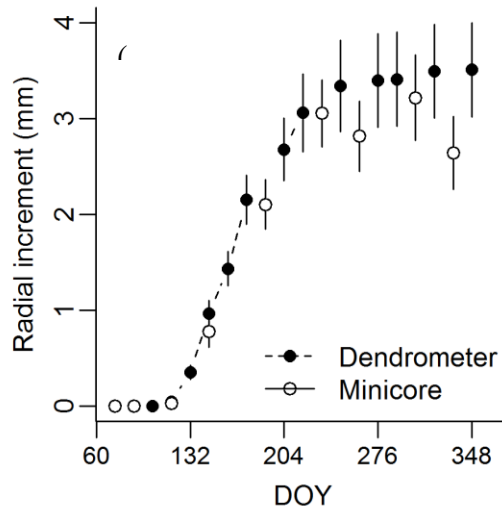
802 Table 1. Likelihood ratio test (LRT) and significance of treatments (thinning intensity,
 803 canopy type, throughfall, month and year) on daily basal area increments of Scots pine
 804 and European beech considering the 2009-2014 period. Bold LRT values indicate
 805 significant effects ($P < 0.05$).

806

Variable	Scots pine		European beech	
	LRT	<i>P</i> value	LRT	<i>P</i> value
Thinning (TH)	1.837	0.39	5.599	0.061
Canopy type (C)	4.056	0.044	Not applicable	Not applicable
Throughfall (T)	860.829	<0.001	69.709	<0.001
Month (M)	1,406.414	<0.001	983.442	<0.001
Year (Y)	139.301	<0.001	99.588	<0.001
TH*C	1.651	0.43	Not applicable	Not applicable
TH*T	6.707	0.034	1.989	0.370
TH*M	33.283	0.006	134.203	<0.001
TH*Y	25.383	0.005	10.16	0.426
C*T	2.552	0.11	Not applicable	Not applicable
C*M	19.864	0.011	Not applicable	Not applicable
C*Y	8.561	0.13	Not applicable	Not applicable
T *M	188.626	<0.001	7.135	0.309
T *Y	120.568	<0.001	32.464	<0.001
TH*C* T	6.059	0.041	Not applicable	Not applicable
TH*C* M	10.851	0.82	Not applicable	Not applicable
TH*C* Y	10.160	0.43	Not applicable	Not applicable
TH* T *M	6.152	0.046	8.179	0.771
TH* T *Y	73.419	<0.001	2.697	0.987
C* T *M	18.659	0.017	Not applicable	Not applicable
C* T *Y	9.715	0.084	Not applicable	Not applicable

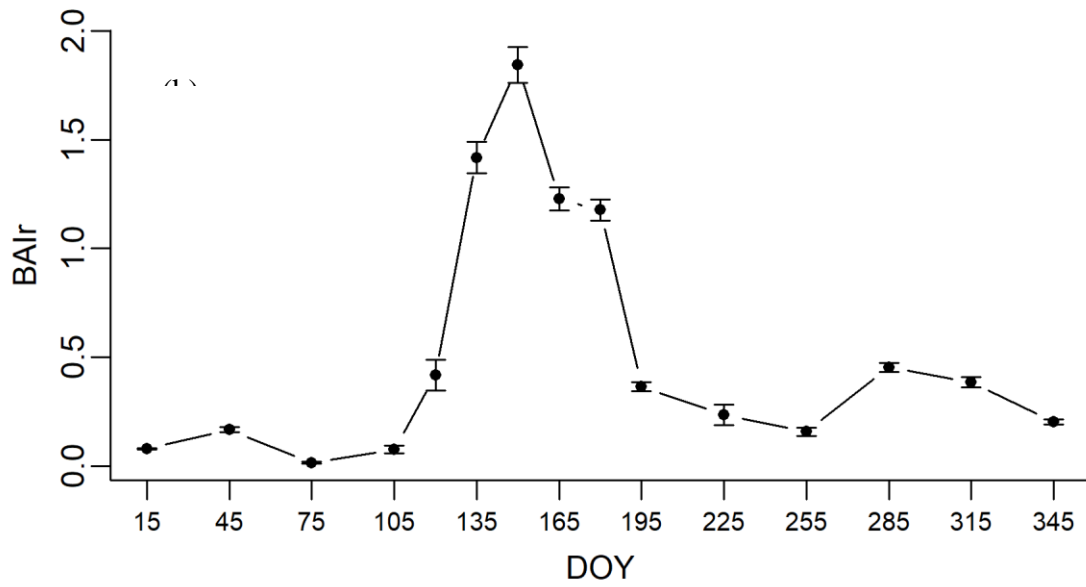
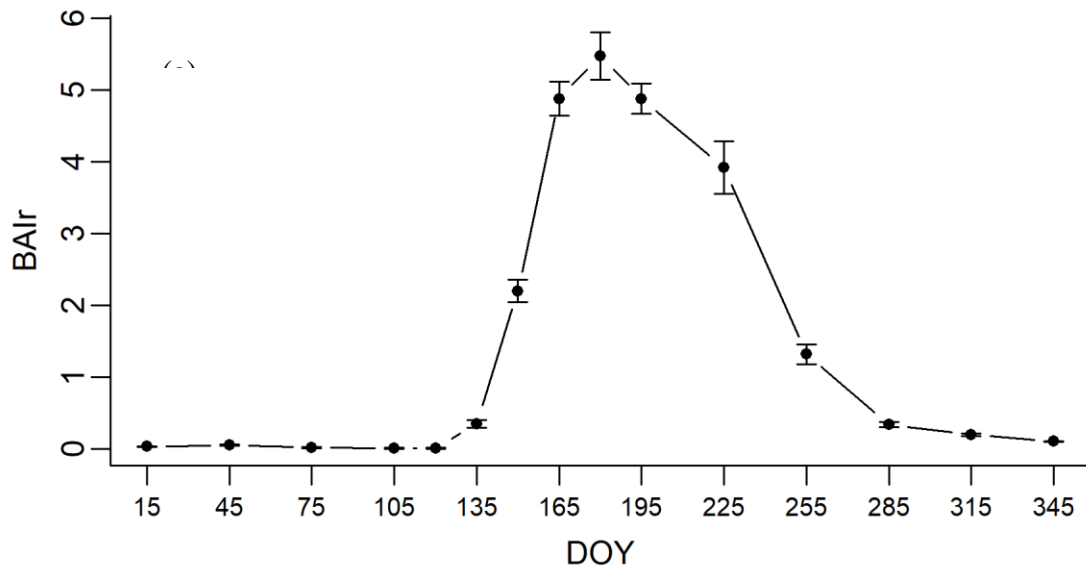
807

808



810

811 Figure 1. Cumulative radial increment for European beech (a) and Scots pine (b) from
812 April to December in 2011 using micro-cores and band dendrometers. DOY stands for
813 day of the year.

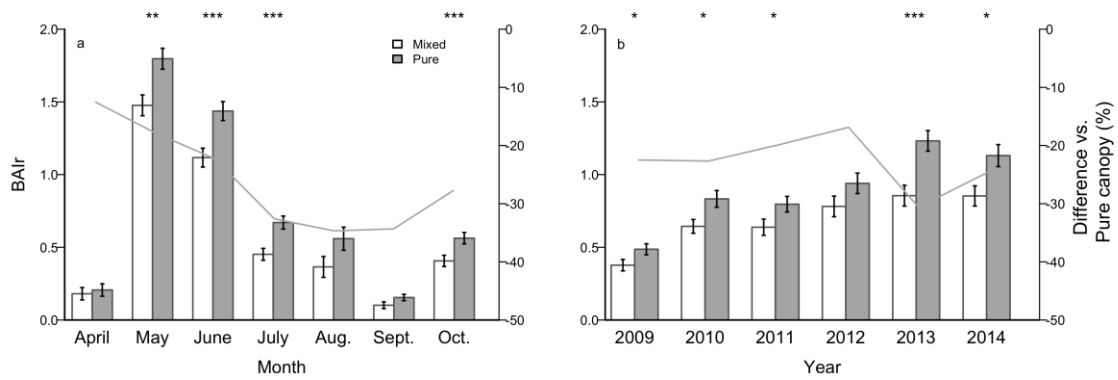


814

815

816 Figure 2. Daily basal area increment (BAIr, cm² day⁻¹; values are means ± SE) for
 817 European beech (a) and Scots pine (b) and from 2009 to 2014. DOY stands for day
 818 of the year.

819



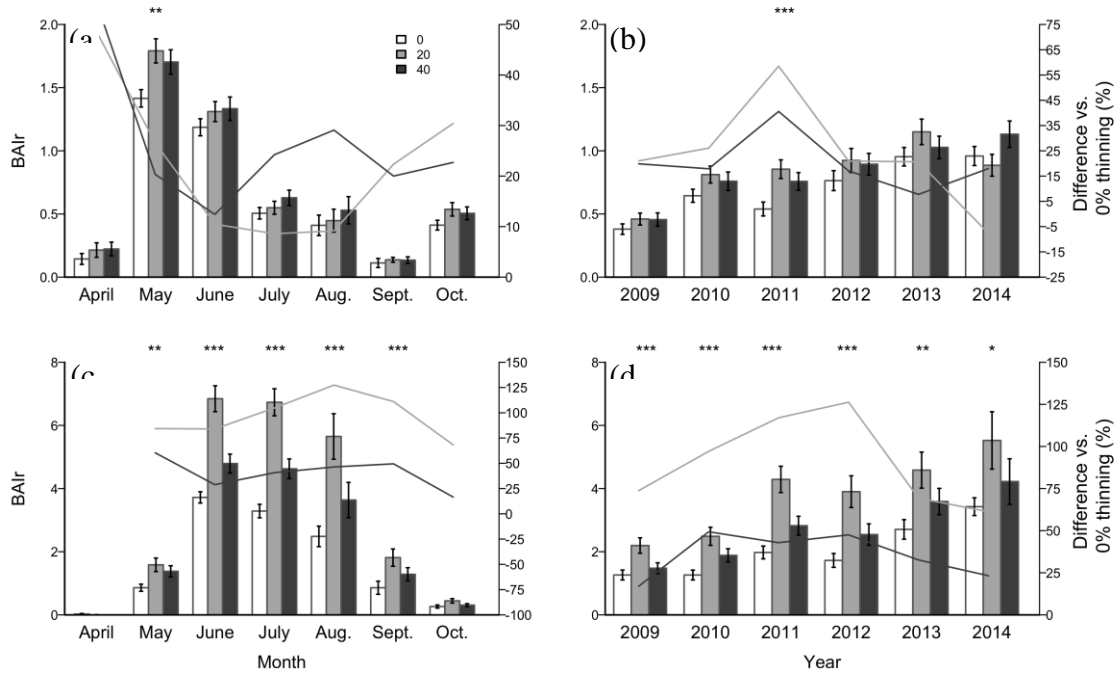
820

821 Figure 3. Monthly (a) and yearly (b) basal area increment rates (BAIr, cm² day⁻¹) of
822 Scots pine as a function of type of canopy (pure pine patches, mixed pine-beech
823 patches) for the growing-season months during the 2009-2014 period. Significant
824 differences in BAIr between types of canopy are indicated as follows: * $P < 0.05$; **
825 $P < 0.01$; *** $P < 0.001$. Percentage pine growth reductions of mixed patches relative to
826 pure patches (light lines) are shown on the right y axis.

827

828

829



830

831 Figure 4. Monthly and yearly basal area increment rates (BAIr, cm² day⁻¹; means ± SE)
 832 of Scots pine (a and b, respectively) and European beech (c and d, respectively) as a
 833 function of thinning intensity (0%, 20% and 40% of basal reduction) for the growing-
 834 season months during the 2009-2014 period. Significant differences in BAIr among
 835 thinning treatments are indicated as follows: * $P < 0.05$; ** $P < 0.01$; *** $P < 0.001$.

836 Percentage growth differences of moderate (20%, light line) and severe (40%, dark line)
 837 thinning relative to unthinned plots are shown on the right y axes.

838

839

840

841

842

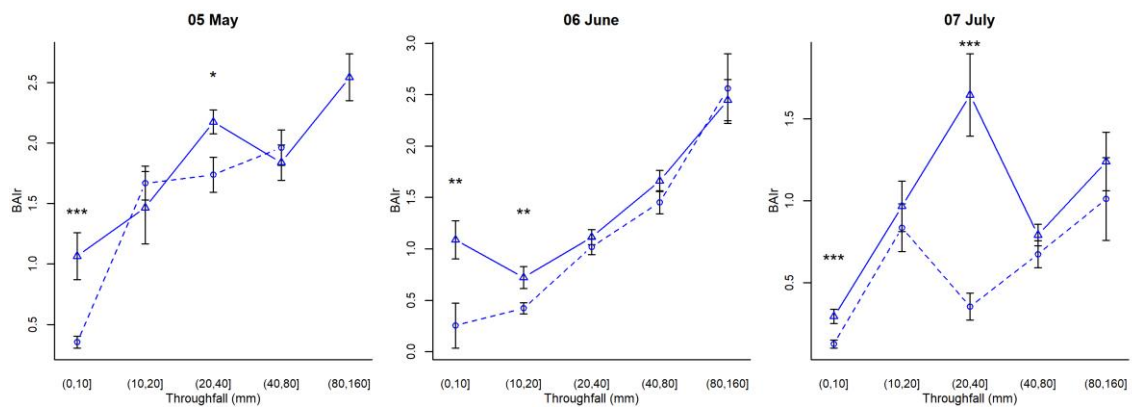
843

844

845

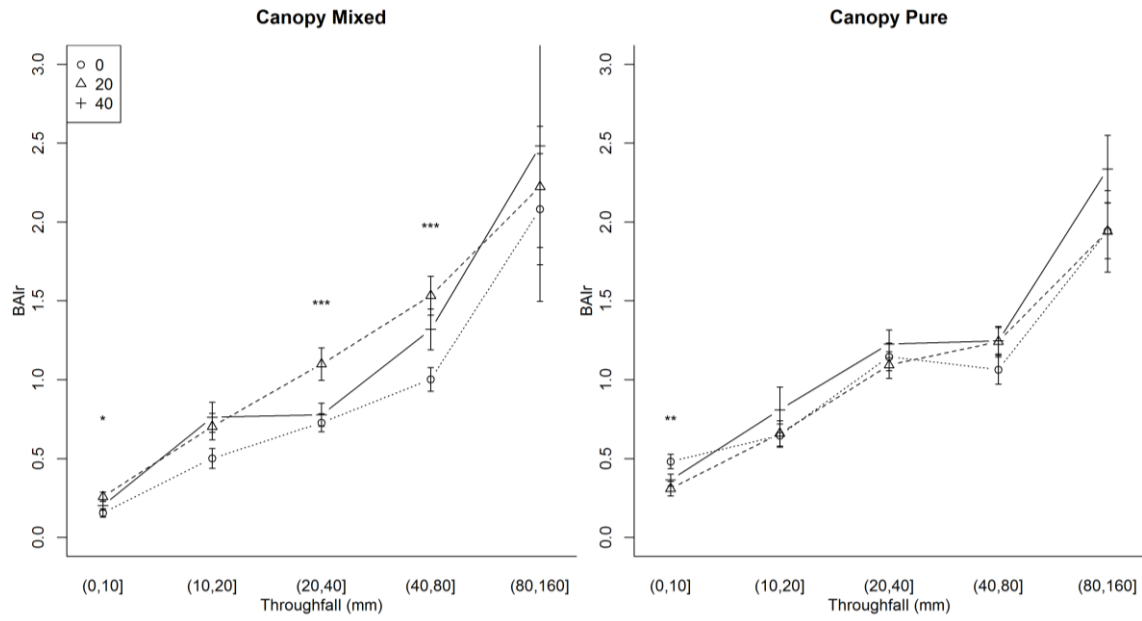
846

847



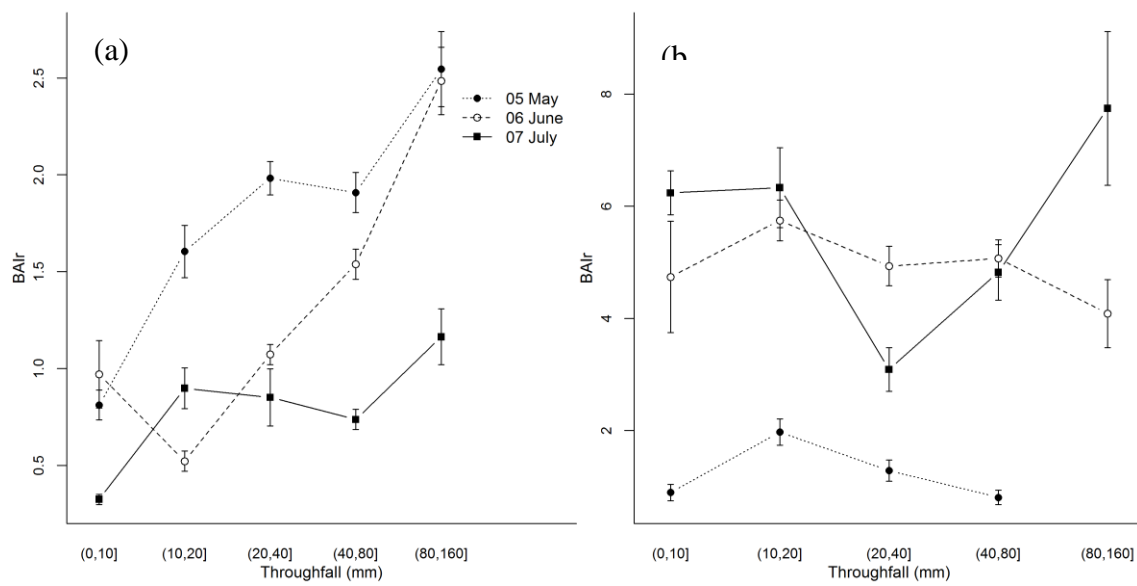
848
 849 Figure 5. Growth rates (daily basal area increment, BAIR in cm² day⁻¹; means ± SE) for
 850 Scots pine in relation to throughfall, canopy type (triangles, pure forest; circles, mixed
 851 forest) during the months with highest tree growth in the 2009-2014 period. Significant
 852 differences in BAIR between canopy treatments are indicated as follows: * $P < 0.05$; **
 853 $P < 0.01$; *** $P < 0.001$. Note the different scales in the “y” axes.

854



855

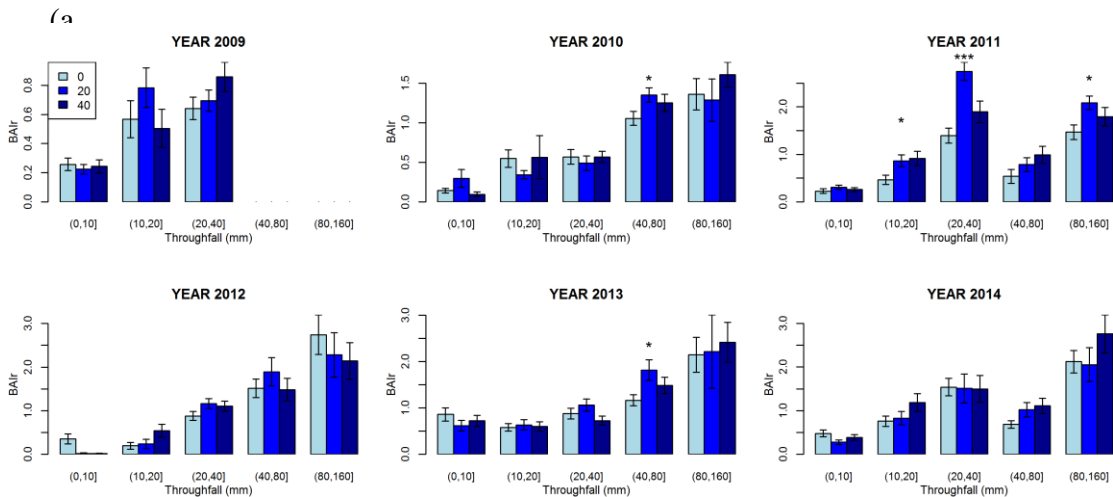
856 Figure 6. Growth rates (daily basal area increment, BAIR, in $\text{cm}^2 \text{ day}^{-1}$; means \pm SE) for
 857 Scots pine in relation to throughfall, thinning (0 %, 20 % and 40 % basal area removed)
 858 and canopy type from April to November and for the 2009-2014 period. Significant
 859 differences among thinning treatments in year month for BAIR are indicated above: *
 860 $P < 0.05$; ** $P < 0.01$; ***



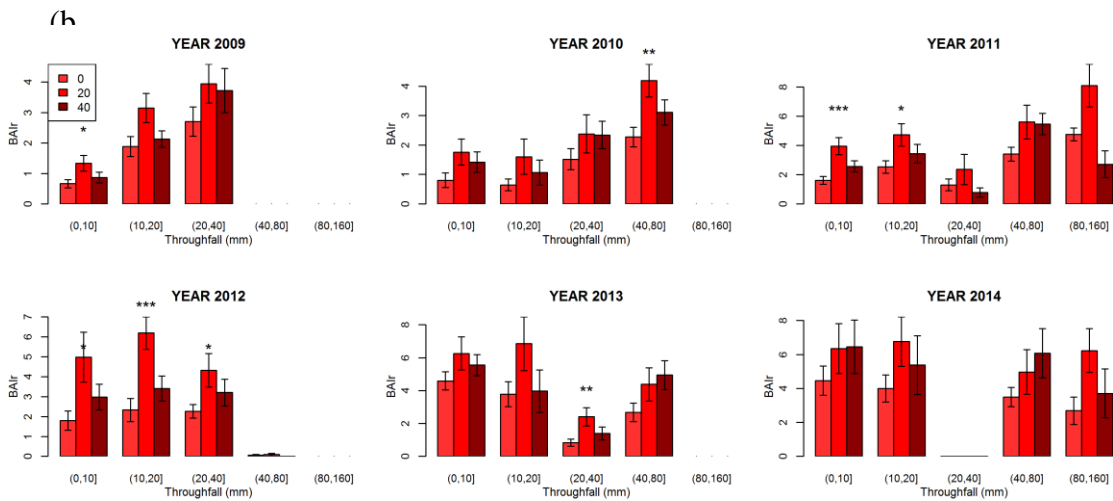
861

862 Figure 7. Growth rates (daily basal area increment, BAIR, in cm² day⁻¹; means ± SE) as a
 863 function of throughfall for Scots pine (a) and European beech (b) during the months
 864 with highest tree growth in the 2009-2014 period.

865



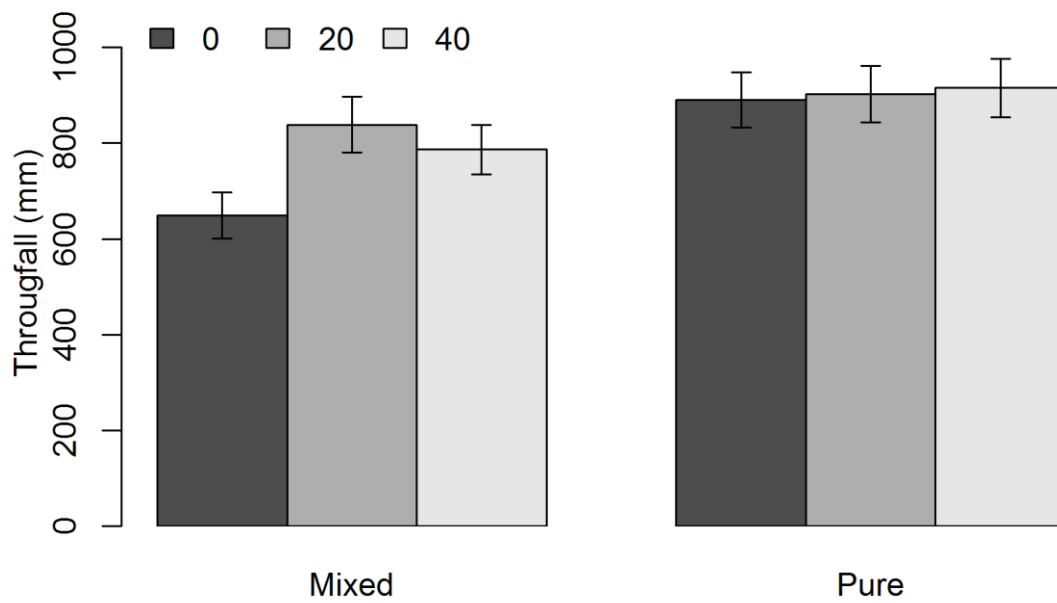
867



868

869 Figure A.1. Growth rates (daily basal area increment, BAIr, in $\text{cm}^2 \text{day}^{-1}$; means \pm SE)
 870 as a function of throughfall and thinning severity (0 %, 20 % and 40 % of basal area
 871 removed) during different years for Scots pine (a; blue bars) and European beech (b; red
 872 bars) considering the 2009-2014 period. Significant differences between thinning
 873 treatments in each year for BAIr are indicated above: * $P < 0.05$; ** $P < 0.01$; ***
 874 $P < 0.001$

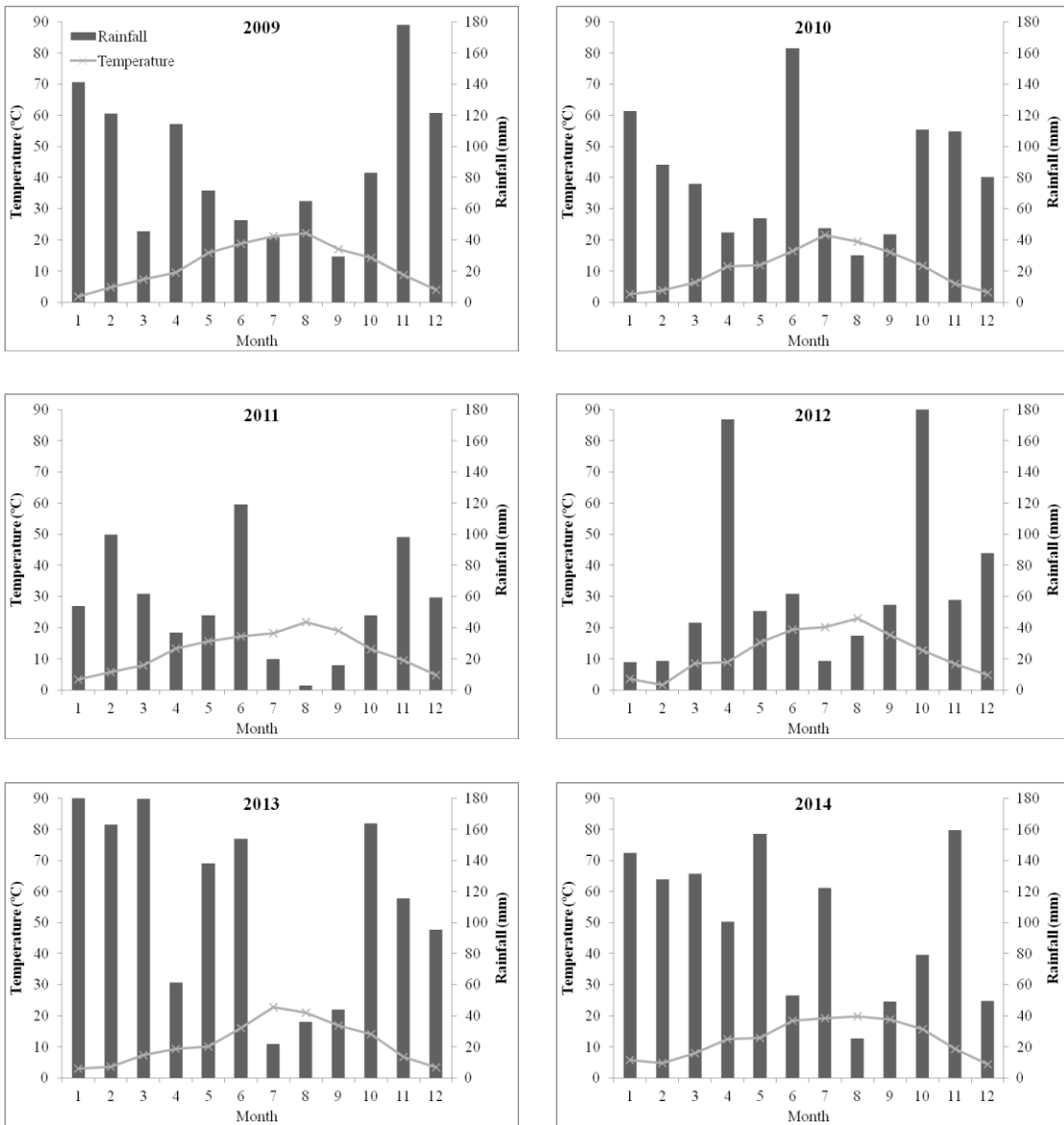
875



876

877 Figure A.2. Mean annual amount of throughfall (means \pm SE) considering thinning
878 intensity (0 %, 20 and 40% of basal area removed) and type of canopy from 2009 to
879 2014.

880



882 Figure A.3. Climographs for the Aspurz study site (Navascués weather station) from
 883 2009 to 2014