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1 **Full title: ASSESSING SOIL PROPERTIES CONTROLLING INTERRILL**
2 **EROSION: AN EMPIRICAL APPROACH UNDER MEDITERRANEAN**
3 **CONDITIONS**

4 Short title: SOIL PROPERTIES CONTROLLING INTERRILL EROSION

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ABSTRACT

14 Soil erodibility is a complex phenomenon that comprises a number of different soil
15 properties. However, most current (empirical) erodibility indices are based on only a few
16 soil properties. A feasible soil characterization of interrill erosion (IE) prediction at large
17 scale should be based on simple, quick, and inexpensive tests to perform. The objective
18 of this work was to identify and assess those soil properties that best reflect soil
19 vulnerability to IE. Twenty-three agricultural soil samples located in Spain and Italy were
20 studied. Forty-nine different physical and chemical soil properties that presumably
21 underpin IE were defined. Experiments were carried out in the field (in microplots using
22 simulated rainfall) and in the lab. The most relevant variables were detected using
23 multivariate analysis. Six key variables were finally identified: RUSLE K factor, a
24 granulometric/organic matter content index, exchange sodium percentage, shear strength,

1 penetration resistance, and permeability of soil seal. The latter is proposed as a useful
2 technique to evaluate soil susceptibility to crusting even when the crust is not present at
3 the time of the field survey. The selected variables represented a wide range of soil
4 properties, and they could also be successfully applied to different soils with different
5 characteristics than those evaluated in our experiments.

6
7 KEY WORDS: interrill erosion, soil erodibility, soil properties, rainfall simulator,
8 multivariate statistical analysis.

9 10 INTRODUCTION

11 Interrill erosion (IE) (Boardman & Poesen, 2006) is a widely recognized form of water
12 soil erosion which recorded rates as high as 49 t ha⁻¹ year⁻¹ (Foster, 1986). The IE involves
13 a relatively uniform removal of soil from the land surface between the rills. The primary
14 force for this erosion is raindrop impact, whose erosive potential depends on raindrop
15 size, distribution, fall velocity, and total mass of impact (Lo, 1992). In addition, IE
16 commonly occurs in areas close to the point of impact of raindrops (within around 1 m)
17 where the detached material could be delivered to nearby rills (if these appear on the
18 landscape) (Bryan, 2000). They can also be deposited downslope by the action of the
19 runoff (Laflen, 2003). During IE, detachment and transport of the soil is principally
20 produced by the impact of raindrops (Lassu *et al.*, 2015; Marzen *et al.*, 2015; Prosdocimi
21 *et al.*, 2016a) and, to a lesser extent, surface runoff. Thus, the IE process implies
22 interaction of diverse factors such as rain, topography, vegetation, stone cover and land
23 management and especially soil erodibility (Gessesse *et al.*, 2015; Gumiere *et al.*, 2009;
24 Keesstra *et al.*, 2014; Ola *et al.*, 2015; Wang *et al.*, 2015). Using a subjective criteria the

1 numerous soil properties related to soil erodibility may be grouped as follows: (1) texture
2 and stoniness of the topsoil (e.g. Dai *et al.*, 2015; Poesen *et al.*, 1994; Poesen and Lavee,
3 1994; Cerdà, 2001); (2) aggregate stability (e.g. Barthès & Roose, 2002; Le Bissonnais,
4 1996; Ma *et al.*, 2014); (3) susceptibility to sealing and crusting (e.g. Arjmand Sajjadi &
5 Mahmoodabadi, 2015; Assouline & Mualem, 2006; Le Bissonnaiss *et al.*, 2005); (4)
6 resistance to shear stress (e.g. Luk & Hamilton, 1986; Léonard & Richard, 2004); and (5)
7 physical-chemical properties (e.g. Ben-Hur & Agassi, 1997; Biswas *et al.*, 2015; Singer
8 *et al.*, 1982).

9 Several indices and parameters have thus been proposed to quantify soil erodibility
10 (Bryan, 2000; Le Bissonnais *et al.*, 2005; Romero *et al.*, 2007; Römken, 1985;
11 Wischmeier & Smith, 1978); however there is not yet a complete understanding of this
12 complex phenomenon. This is why most of the current models to estimate soil erodibility
13 are still empirical based on statistical relationships between inputs (soil properties) and
14 outputs (erosion rate). Despite the large number of variables intervening in the erosion
15 process, most of the current empirical indices are only based on a limited number of
16 variables –sometimes only one or two– that are assumed *a priori* to be dominant in the
17 erosion process under the study conditions. On the other hand, evaluation of large-scale
18 soil erodibility, e.g. catchment scale, should consider variables that are obtained from
19 simple and economical procedures and techniques (Le Bissonnais *et al.*, 2005).

20 Commonly, the relationships between erodibility and soil properties have been
21 explored using classic statistical procedures like correlation matrix, simple linear
22 regressions and multivariate techniques (e.g. Arnaez *et al.*, 2007; Cerdan *et al.*, 2010;
23 García-Ruíz *et al.*, 2015; Maetens *et al.*, 2012; Prosdocimi *et al.*, 2016b; Sheridan *et al.*,
24 2000; Verhaegen, 1984).

1 The aim of this research was to identify and assess soil properties that are easily and
2 quickly determined and that best reflect the soil vulnerability to interrill erosion.

3 The experimentation was conducted on agricultural soils of Navarre, León (Spain) and
4 Sicily (Italy). Experiments were carried out in microplots with simulated rainfall and in
5 the laboratory. Forty nine soil variables and parameters were obtained via techniques and
6 methodologies proposed in the literature. Data were analyzed using multivariate statistics
7 tools.

8 9 MATERIALS AND METHODS

10 *Study areas*

11 Twenty-three agricultural soils located in three studied areas: (1) León (Spain); (2)
12 Navarre (Spain); and (3) Sicily (Italy) were assessed (Figure 1 and Data S1, Google Earth
13 file). These studied areas are commonly affected by water erosion (e.g. Casalí *et al.*, 1999;
14 Casalí *et al.*, 2008; Capra *et al.*, 2012). The main crop in all studied areas was winter
15 wheat with the exception of one situation (LEO 1; Table I), where the crop was rye. In
16 León study area (NW Spain), with an annual rainfall average of 449 mm and a
17 Mediterranean climate, 2 soils were located: LEO 1, and LEO 2 (Table I). Under a
18 Mediterranean climate and with a mean precipitation range between 547 and 1310 mm,
19 11 soils were located in the study area of Navarre (N Spain): PIT 1 to PIT 3, AOI 1 to
20 AOI 6, LUM 1 to LUM 3, ABA 1, and ABA 2 (Table I). Finally, 7 soils were located in
21 the Sicilia (S Italy) study area: RAD 1 to RAD 7 (Table I). This last area is characterized
22 by a Mediterranean climate type with an average annual rainfall of 500 mm

1 The experimentation period was from October to March in the years 2012, 2013 and
2 2014. In this period the soil moisture is high and almost constant, due to the accumulation
3 of ca. 80% of the annual precipitation.

4 *Experimentation protocol*

5 Rainfall simulation experiments with a drip portable rainfall simulator (Eijkelkamp
6 Agrisearch Equipment, model 09.06, the Netherlands) designed by Kamphorst (1987)
7 were performed at each of the 23 soils studied. This type of rainfall simulator has been
8 widely used to assess soil erodibility (e.g. Bagarello & Ferro, 2004; Iserloh *et al.*, 2013;
9 Martínez-Zavala & Jordán, 2008; Romero *et al.*, 2007). It consists in two main parts: (1)
10 a capillary sprinkler for control the rain shower produced by 49 tubes located at a height
11 of 0.4 m from the soil surface; and (2) a partly buried square metal frame which delimited
12 the test plot (0.0625 m²) and also prevented the lateral movement of water (Figure 2). In
13 all studied areas the experimental microplot was located in points of the landscape which
14 had a similar terrain slope and where the IE phenomenon were evident.

15 Three replicate rainfall simulations were conducted at each selected site. The mean
16 intensity of the rain was 131.5 mm h⁻¹ and the kinetic energy of drops was 4 J m⁻² mm⁻¹.
17 The estimated mean annual 30 min maximum rainfall intensity for 10 years is 71.24 mm
18 h⁻¹ for the studied areas. This high intensity was adopted to correct the lower kinetic
19 energy of the rain shower generated by the rainfall simulator due to its low fall height
20 (Iserloh *et al.*, 2013).

21 Rainfall simulations were continued as long as necessary to reach a roughly constant
22 runoff rate, i.e. soil moisture around saturation. This steady flow was achieved in the
23 different experiments after around 20-30 minutes. From this moment, water and sediment
24 samples were collected every 2 minutes during approximately 20 minutes. These were

1 dried in a stove (105 °C, 24 h) to obtain the weight of eroded soil. Then, the erosion rate
2 ($\text{t ha}^{-1} \text{h}^{-1}$) (SL_1 , Table II) was calculated (e.g. Cerda et al., 2016; Prodoscimi et al., 2016a;
3 Rodrigo Comino et al., 2016a). The basic infiltration rate (HY_1 : mm h^{-1} , Table II) was
4 calculated as the difference between the volume of rain and that of stabilized runoff.

5 Various soil variables highlighted in the literature as soil erodibility potential drivers
6 (a total of 49 described in Table S1) were determined in each study situation. Briefly, the
7 different methodologies carried out are described below (see Table S1 for more
8 information). A composed soil sample from the first 15 cm of soil depth were collected
9 from various points adjacent to the microplot on each of the 23 studied soils. After oven-
10 drying and sieving ($< 2\text{mm}$), different determinations were performed in the laboratory
11 in order to determinate a large set of different soil parameters: (1) aggregate stability to
12 the 3 disaggregation methods proposed by Le Bissonnaiss (1996): slaking (SI_4), clay
13 swelling (SI_5), and mechanical breakdown by shaking (SI_6); (2) crusting susceptibility
14 index C_{5-10} (CR_1) (De Ploey & Múcher, 1981); (3) relative sealing index (CR_2) (Pla,
15 1982); (4) seal hydraulic conductivity (HY_2) (Pla, 1982); (5) 4 crusting indices obtained
16 from soil texture and organic matter (CR_3 , CR_4 , CR_5 , and CR_6) (Comerma et al., 1992;
17 FAO, 1980; Florentino, 1998); (6) various physical (i.e. texture, GT_{1-5} and GF_{1-5} ;
18 stoniness, GC_{1-6}), chemical (i.e. organic matter percentage, CH_1 ; electrical conductivity,
19 CH_2 ; exchangeable sodium percentage, CH_3 ; pH of soil saturated paste, CH_4 ; cation
20 exchange capacity, CH_5 ; and calcium carbonate, CH_6) and physical-chemical soil
21 variables (i.e. 3 structure stability indices, SI_1 , SI_2 , SI_3 ; K factor of RUSLE, E_2); and (7)
22 critical shearing (SS_5) and soil erodibility coefficient (E_1) determined by a jet test
23 apparatus by using undisturbed soil samples (Hanson & Cook, 2004) (more details in
24 Table S1, L). Additionally, several soil properties were directly determined on the field.

1 This was done both on the microplot itself (Table S1, F_{IN}) and at sites near to it (Table
2 S1, F_{OUT}). The following parameters were measured within the microplot: (1) hydraulic
3 conductivity of the crust formed after the drying of the soil seal produced by the rain
4 simulation (HY_3) (Boiffin & Monnier, 1985); and (2) 3 variables of crust resistance to
5 penetration after the rain shower (PR_1 , PR_2 , and PR_3) (Bradford et al., 1992; Truman &
6 Bradford, 1990). And out of the microplot the following ones: (1) bulk density (PH_1) and
7 (2) 4 different soil shear strength measurements (SS_1 , SS_2 , SS_3 , and SS_4) (Léonard &
8 Richard, 2004).

9 *Statistical analyses*

10 The mean value of the variables was further considered for statistical analysis (see
11 Table S1). The relationships between the IE rate and the rest of the variables were
12 investigated using 3 different multivariable statistical tools: Cluster Analysis (CA),
13 Principal Components Analysis (PCA) and Multiple Regression Analysis (MRA). The
14 results obtained with each statistical tool were analyzed independently from each other.
15 The suitability of the dataset for applying those statistical techniques was assessed via
16 Bartlett's sphericity test (Bartlett, 1954). All analyses were performed employing version
17 3.1.1 of the R statistics software (R Core Team, 2015).

18 The CA is a non-supervised classification technique for the recognition of similar
19 behavior patterns between observations (Vega *et al.*, 1998). This technique uses all the
20 information supplied by the original data set without making any previous assumptions.
21 In this study, a hierarchical agglomerative CA was proposed on the data via Ward's
22 method (Ward, 1963) and using the Euclidean distance as a similarity measurement. This
23 method was selected for its greater potential as a tool for the creation of groups (Willet,
24 1987). Thus, each identified group or cluster is characterized by mean values of the

1 variables defining that cluster (mean in cluster), which are significantly different from the
2 respective mean value of the entire population (overall mean) (e.g. Anderberg, 1973;
3 Kabacoff, 2015).

4 The PCA reduces the dimensions of large-size datasets to restrict the number of
5 explanatory variables without losing important information in the process (Shrestha &
6 Kazama, 2007). On some occasions, it is necessary to orientate the PCA towards those
7 variables of special interest (supplementary variables) and therefore to refer the analytical
8 process to them (Giménez *et al.*, 2012). In this study, the supplementary variable selected
9 was SL_1 (Table S1). Thus, PCA transforms the original variables into new non-correlated
10 synthetic variables called Principal Components (PCs). These PCs are useful instruments
11 to study possible relationships and to identify trends between those variables with higher
12 correlation to PCs. When the correlation values between the PCs and the variables are too
13 high (or low) it may be difficult to interpret the results, but rotating the PCs can solve this
14 problem (Westra *et al.*, 2010). In this study, the Varimax rotation criterion (Kaiser, 1958)
15 was selected and applied to the PCA.

16 The MRA allow obtaining the best linear relation between the variables analyzed
17 relating a selected dependent variable (SL_1) as a function of a set of independent variables.
18 In this study, all the possible linear models from one to four of the soil variables were
19 obtained and analyzed to seek the best explanatory model (i.e. principle of parsimony,
20 Vandekerckhove *et al.*, 2015). In each of the resulting models, the following evaluation
21 criteria were applied to diagnose the best regression model: (1) calculation of the variance
22 inflation factor (VIF, equation 1) to discard the independent variables presenting
23 multicollinearity (i.e. $VIF > 2$) (Lin, 2008); (2) outliers were discarded when the
24 standardized error in the regression residues was larger or smaller than 2 (Kabacoff,

2015); (3) Akaike information criterion (AIC, equation 2) values as a scale to select the model's predictors (Akaike, 1974); (4) verifying the goodness of fit obtained by the model employing the Nash-Sutcliffe efficiency coefficient (NSE, equation 3) (Nash & Sutcliffe, 1970), the mean squared error (MSE, equation 4) and the root mean squared error (RMSE, equation 5) (Moriassi *et al.*, 2007); and (5) regression diagnosing based on the significance level ($p < 0.05$, t-test) for the regression coefficients of each independent variable conforming the model (Walpole *et al.*, 2011).

$$VIF = \frac{1}{1-R^2} \quad (1)$$

$$AIC = 2k - 2\ln(L) \quad (2)$$

$$NSE = 1 - \frac{\sum_{i=1}^n (S_i - O_i)^2}{\sum_{i=1}^n (O_i - \bar{O})^2} \quad (3)$$

$$MSE = \frac{1}{n} \sum_{i=1}^n (O_i - S_i)^2 \quad (4)$$

$$RMSE = \sqrt{\frac{1}{n} \sum_{i=1}^n (O_i - S_i)^2} \quad (5)$$

where k is the number of parameters included by the model; L is the maximized value of the model's verisimilitude function; S_i is the simulated value for the i case; O_i is the value observed for the i case; \bar{O} is the mean value for the observed cases; n is the number of observations; and R^2 is the multiple determination coefficient of the fitted model.

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RESULTS

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The Bartlett test showed a probability $p < 0.05$ (significance at 95% level), which – together with a value of the correlation matrix determinant equal to 0 and with the existence of correlation coefficients between variables of above 0.30 in the same matrix (not shown)– suggests that the variables have a sufficient level of correlation to use the 3 multivariate statistics tools proposed here.

1 *Cluster analysis (CA)*

2 The CA defined the existence of 2 clusters (Cluster 1 and Cluster 2). Each of the
3 clusters was constituted by approximately half of the soils: 11 in Cluster 1 and 12 in
4 Cluster 2. The soils grouped in Cluster 1 presented a 3-fold higher mean erosion rate (SL_1)
5 than that corresponding to the soils in Cluster 2 (median of $6.6 \text{ t ha}^{-1} \text{ h}^{-1}$ and $2.4 \text{ t ha}^{-1} \text{ h}^{-1}$,
6 respectively). This result suggests that the variables in their respective value ranges
7 discriminated by the CA would be related to some extent to the erodibility of the soil. Of
8 the 19 variables identified (Table II), only those with the highest statistical weight were
9 selected; i.e. those showing a significance level of $p < 0.005$ in the Student t-test
10 (evaluating that the mean of the variables in each cluster differed statistically from the
11 mean value of the whole population): CH_5 , CR_5 , SI_2 , SI_3 , E_2 , CR_4 , CR_3 , and PR_3 . These
12 variables were then grouped according to its typology in the following groups: (i) cation
13 exchange capacity (CH_5), (ii) soil mechanical resistance to penetration in the first 6 cm
14 (PR_3) and (iii) variables obtained basically via a balance between the soil texture (particle
15 size $< 2 \text{ mm}$) and the organic matter present in the soil (CR_5 , SI_2 , SI_3 , E_2 , CR_4 , and CR_3).
16 As these last variables were similar with regard to their nature and statistical weight, E_2
17 was selected as being representative of this last group due to its wide support as a soil
18 erodibility parameter in the literature (e.g. Meyer & Harmon, 1984, Romero *et al.*, 2007,
19 Wischmeier & Smith, 1978).

20 However, variable CH_5 provided ambiguous information on soil erodibility as
21 mentioned later in the discussion section and was rejected. Then, the variables resulting
22 from the CA were E_2 and PR_3 .

23 *Principal Components Analysis (PCA)*

1 After fixing SL_1 as the supplementary or reference variable those PCs presenting a
2 lower eigenvalue than unity after their rotation were discarded (Andrews *et al.*, 2002).
3 This resulted in 11 PCs (Table III). The first 2 PCs (PC_1 and PC_2) were capable of
4 explaining 38.4% of the total variance of the data (21.7% and 16.8%, respectively), but
5 PC_2 was discarded due to its lower correlation with variable SL_1 (0.19); PC_1 had a
6 correlation of 0.35. In addition, only those variables from PC_1 with a factor correlation \geq
7 0.5 were further considered for analysis (Table III).

8 Thus, 13 variables were selected and divided according to their type into 5 groups: (i)
9 composition of the soil texture and organic matter content (SI_1 , SI_2 , SI_3 , CR_5 , CR_6 , and
10 E_2); (ii) soil resistance to penetration in the first centimeters of depth (PR_2 , PR_3 , and PH_1);
11 (iii) soil shear strength (SS_1); (iv) soil chemical composition (CH_2 and CH_3); and (v)
12 permeability of the soil crust measured in the laboratory (HY_2). Of the groups showing
13 more than 1 variable (i, ii and iv), the following were selected: (1) E_2 due to its higher
14 correlation value within PC_1 and therefore with SL_1 (0.86 in absolute value, Table III); (2)
15 PR_3 with a higher statistical weight although similar to that of PR_2 (0.51 in absolute value,
16 Table III) and also for having been prominent in the CA (see above); and (3) CH_3
17 assuming that the high electrical conductivity values (CH_2) were precisely a result of the
18 high exchangeable sodium percentage (CH_3). From the PCA, the outstanding variables
19 were E_2 , PR_3 , SS_1 , CH_3 , and HY_2 .

20 *Multiple regression analysis (MRA)*

21 Similarly, the MRA weighted those variables of the total population in the study (Table
22 S1) that would best fit an explanatory linear regression model of SL_1 . All possible models
23 with one variable (48), two variables (1128), three variables (17296) and four variables
24 (194580) were obtained.

1 The best model with one variable (Model 1) had CH₃ as an independent variable. This
2 model showed the highest value of NSE (0.92), the lowest MSE (0.04), RMSE (0.21) and
3 AIC (3.18) from all models with one variable. With respect to the best model with two
4 variables (Model 2), this kept CH₃ as the first variable and introduced CR₆ as the second
5 independent variable. Model 2 showed a lower value for AIC than Model 1 (-9.3 and 3.18,
6 respectively, Table IV). It should be remembered that high values for AIC indicate a loss
7 of quality in the goodness of the model caused by the complexity of the model itself and
8 its goodness of fit (Akaike, 1974). Models with more than 2 explanatory variables were
9 discarded due to their low accuracy (i.e., higher AIC values). Therefore, Model 2
10 (equation 6) was selected as having the best relationship between soil properties (CH₃ and
11 CR₆) and the IE rate (SL₁).

$$12 \quad SL_1 = 0.813 + 0.151 \cdot CH_3 + 0.575 \cdot CR_6 \quad (6)$$

13 Both variables CH₃ and CR₆ already stood out in the PCA (Table III); only E₂ was
14 selected instead of CR₆ for the reasons given above.

15 *Guide values of the variables selected*

16 The overall evaluation of the 3 statistical analyses identified 6 key soil variables in the
17 establishment of soil erodibility under the experimental conditions of the study: E₂ (K
18 factor of RUSLE), CH₃ (exchangeable sodium percentage), PR₃ (penetration resistance
19 in the first 6 centimeters of the soil depth), SS₁ (shear strength), CR₆ (modified crusting
20 index), and HY₂ (hydraulic conductivity of seal) (see Table S1). Moreover, a guide value
21 is determined from the mean value of each of the aforementioned six variables reaches in
22 the least susceptible cluster (Cluster 2, Table V). Thus, the transition between erosion-
23 resistant and vulnerable soils (Cluster 1, Table V) can be roughly defined.

1 Figure 3 shows the distribution of the mean values and the standard deviations for
2 these 6 key variables in the 2 clusters. In this way, a new soil (with a similar texture and
3 structure to those presented here) could be roughly classified as most resistant to erosion
4 if it shows values of approximately $0.02 \text{ t ha h ha}^{-1} \text{ MJ}^{-1} \text{ mm}^{-1}$ for E_2 , 0.75 for CR_6 , 4.19%
5 for CH_3 , 500 kPa for PR_3 , 16.00 kPa for SS_1 and 2.60 mm h^{-1} for HY_2 (see Figure 3 and
6 Table V). These guide values must be interpreted independently from one another because
7 we do not evaluate the possible interaction among the proposed key variables.

8

9

DISCUSSION

10 First of all, it is important to highlight that the individual relations between the IE rate
11 and a group of soil properties were established through statistical analysis but without
12 examining interdependency among those variables.

13 The susceptibility to erosion in the soils analyzed was due to the action of diverse
14 factors. The first factor is the low content in clay and organic particles, both well-known
15 cementing agents of soil aggregates (Boix-Fayos *et al.*, 2001). For example, Wilcox &
16 Wood (1989) applied a rain simulation on 88 microplots of loam and clay soils in New
17 México (USA) and showed that IE rates (between 0.03 and 8 t ha^{-1}) correlated negatively
18 with the increase in clay content, although the authors did not define any threshold value
19 for this relationship. Similar results were obtained by Meyer & Harmon (1984) who
20 recorded the greatest resistance to IE in 18 agricultural soils with contrasting textures.
21 This was determined by rain simulation in 7 soils in which the clay content exceeded
22 27%. These results are comparable to those obtained in our study in which those soils
23 least vulnerable to erosion (Cluster 2) showed a clay content of approximately 37% (not
24 shown). On the other hand, the effect of organic matter on the improvement of the

1 structure stability of the soil –and thus resistance to erosion– is extensively documented
2 (e.g. Barthès & Roose, 2002; Bryan, 2000; Cerdan *et al.*,2010; Dimoyiannis *et al.*, 1996;
3 Sheridan *et al.*, 2000; Prosdocimi *et al.*,2016b). Le Bissonnais & Arrouays (1997) studied
4 loam soils in France and established that 2.5 to 3.0% of organic matter improved and
5 maintained the stability of their aggregates against the action of simulated rain. This
6 reduced the susceptibility to crusting and erosion. In our experiments, the soils most
7 resistant to erosion presented organic matter values very close to the threshold cited above
8 (Table II). In this line, Rodrigo Comino et al. (2016c), comparing the variability of soil
9 erosion in Mediterranean vineyards in Spain, registered the highest values of soil loss
10 (generated by rainfall experiments) in soils with both the lowest organic matter content
11 and the highest concentration of silt particles.

12 In fact, E_2 and CR_6 are important explanatory variables for soil erodibility because
13 both are ultimately defined as a balance between the soil texture and organic matter (Table
14 S1). Meyer & Harmon (1984) and Panagos et al. (2014) –the latter from an analysis of
15 approximately 20.000 soil samples obtained from 25 European countries– defined a
16 critical value of E_2 of 0.030-0.035 $ta h MJ^{-1} ha^{-1} mm^{-1}$ to classify soils as being most
17 susceptible to erosion. This threshold is very close to that obtained in our studies (Table
18 V). FAO studies (1980) showed that soils with a CR_6 value over 2 are highly susceptible
19 to crusting, although Pulido-Moncada et al. (2009) reduced this value to 0.70 in a study
20 of 5 agricultural soils in Venezuela. The latter agrees fairly well with the mean value
21 determined in this study ($CR_6 = 0.74$) for the most erodible soils (Cluster 1). However,
22 the most resistant soils (Cluster 2) gave a mean CR_6 value of only slightly over 1. It should
23 be noted that E_2 , unlike CR_6 , also includes a permeability and soil structure class
24 quantification (Table S1). While both of these behaved similarly as explanatory variables

1 for soil erodibility, in more contrasting soils –especially with regard to texture and
2 structure to those of our experiments– E₂ would probably stand above CR₆.

3 Second, the structure stability of some soils –and consequently their vulnerability to
4 erosion– was affected by a high content of Na⁺ cations in the soil exchange complex. It
5 is known that high sodicity can lead to an important physical degradation of the soils
6 (Ben-Hur & Agassi, 1997; Shainberg & Letey, 1984). This increases susceptibility to
7 sealing and eventually to erosion. For example, Mamedov et al. (2002) evaluated the
8 effect of different exchangeable sodium percentages (ESP) (2, 5, 10 and 20%) and clay
9 contents (between 22 and 62%) on soil losses triggered by simulated rain in 6 arable
10 Israeli soils. These authors determined that the erosion rate increased as the ESP
11 augmented and the clay content in the soils diminished. Other laboratory studies with
12 simulated rain by Singer et al. (1982) showed that 7 loam, silty loam and silty clay loam
13 soils had an increase of over 20% in the IE rate when the ESP value increased from 4 to
14 54% as a result of the negative effect of the Na⁺ on the resistance of soil aggregates to
15 erosion. Kemper & Koch (1966) confirmed this in a study of 519 soils in the USA and
16 Canada. The lowest values of aggregate stability were observed when the ESP value was
17 over 20% and the clay content was below 22%. The latter trends were similar to those
18 sodic soils identified in Cluster 1 (PIT 1, PIT 2, RAD 5, and RAD 7). These 4 soils
19 exhibited on average a higher ESP value (49.25%, not shown) and a lower clay content
20 (25.38%, not shown) with respect to the rest of the soils analyzed (4.36% and 31.25%,
21 respectively, and not shown). While the cation exchange capacity was another notable
22 soil property in our analysis (variable CH₅, see above), it did not have any effect *per se*
23 on the erosion but would be conditioned by the type of cations dominating in the soil
24 exchange complex. Thus, if bivalent cations dominate, then the soil aggregation would

1 benefit. But if the dominant cation is sodium (see above), then there is an opposite effect
2 (peptization) (Bronick & Lal, 2005). Hence, we decided to exclude the variable CH₅ from
3 the key explanatory variables selected as discussed in the previous section.

4 Third, soil susceptibility to sealing is also important and it was quantified via
5 permeability values of the seal formed in the laboratory after rain simulation (variable
6 HY₂). Thus, lower values of HY₂ would lead to higher runoff and erosion rates in the field
7 (Ben-Hur & Agassi, 1997). Ramos et al. (2003) proposed a minimum threshold value of
8 5 mm h⁻¹ for the variable HY₂ under which the risk of sealing and erosion in soils
9 cultivated with cereal and vineyards in Catalonia (Spain) was increased. Those authors
10 related the increase in the value of HY₂ to a reduction in the percentage of medium-sized
11 particles (fine sands and silts) and to a higher content of organic matter in the soil. In our
12 experiments, both Clusters 1 and 2 soil groups showed a lower value of HY₂ than that
13 proposed by Ramos et al. (2003): 1.84 and 2.73 mm h⁻¹, respectively. Even so, we can
14 claim that the crusting problem plays a key role in the soil erodibility of the soils assessed
15 based on field observations. This statement has been also supported by numerous research
16 carried out on cultivated soils of Spain and Italy (e.g. Cerdà, 1997; Cerdà *et al.*, 2016;
17 Marzen *et al.*, 2015; Prosdocimi *et al.*, 2016b; Ramos & Martínez-Casasnovas, 2007;
18 Rodrigo Comino *et al.*, 2016b,c). On the other hand, while variable HY₂ was originally
19 proposed to detect soil sealing susceptibility (Pla, 1982), it also supplies information on
20 aggregate stability because the seal is formed with the soil particles disaggregated due to
21 action of rainfall (Assouline & Mualem, 2006). As a matter of fact, none of the 3 variables
22 (SI₄, SI₅, and SI₆) proposed by Le Bissonnais (1996) to evaluate structure stability stood
23 out in our assessment. This was unexpected because several works have shown a clear
24 relationship between one or other of these variables with crusting and subsequent erosion

1 of the soil under simulated rain (e.g. Amézqueta *et al.*, 1996; Barthès & Roose, 2002; Le
2 Bissonnais, 1996; Leguédouis & Le Bissonnais, 2004; Ma *et al.*, 2014; Ramos *et al.*, 2003;
3 Wang *et al.*, 2015). Thus, the disaggregation and subsequent sealing of our soils was a
4 direct result of the impact of the raindrops (variable HY₂, see above); this erosive agent
5 was not properly reflected in the variable SI₆ (Le Bissonnais, 1996) in which the action
6 of the rain try to be simulated by agitating the soil aggregates. Thus, HY₂ would reflect
7 both the susceptibility of the soil to sealing and its aggregate stability. It is important to
8 use a technique that evaluates the soil potential to crusting without the crust being present
9 in the field at the moment of determination.

10 Finally, erosion susceptibility was somewhat reflected in the soil resistance to shear
11 strength and/or to penetration (variables SS₁ and PR₃, respectively). Concretely, increases
12 in the value of SS₁ would be related to a greater soil resistance to detachment caused by
13 raindrop impact and surface runoff (Léonard & Richard, 2004). Meanwhile, high values
14 of PR₃ would be associated with soil physical degradation from rain. This led to soil
15 compaction and, consequently, a diminution in its permeability (Herrick & Jones, 2002).
16 That is why our mean values of PR₃ and SS₁ were 50% lower and 15% higher,
17 respectively, in the soils most resistant to erosion (Cluster 2) than in the more vulnerable
18 ones (Cluster 1). Luk & Hamilton (1986) studied 2 non-agricultural soils (loam and silty
19 loam) and proposed a shear strength threshold value (measured with a vane shear
20 apparatus) between 13 and 15 kPa. Below this, the IE rate –determined by in-field rain
21 simulation– increased. In our experiments, the least erodible soils (Cluster 2) gave a
22 higher SS₁ value (16.00 kPa, Table V) than the threshold cited. However, the soils most
23 susceptible to erosion (Cluster 1), had a value of this variable that was lower (14.11 kPa,
24 Table V). However, this was within the previous range and these data should be

1 interpreted with caution because Luk & Hamilton (1986) studied non-agricultural soils
2 and our study used agricultural ones. It is worth noting that the variable SS₁ was obtained
3 with a disk of our own design with a high depth/diameter ratio (Table S1), in order to
4 increase the volume of the soil submitted to shearing and thus obtained more
5 representative measurements of this parameter (Léonard & Richard, 2004). For the
6 variable PR₃, mean values almost twice larger were obtained in Cluster 1 (916.15 kPa)
7 than in Cluster 2 (462.46 kPa). Although several works report general trends on the
8 compaction effect on the soil after a rain event and the subsequent variation on the IE rate
9 (e.g. Laufer *et al.*, 2016; Neave & Rayburg, 2007; Truman & Bradford, 1990), no clear
10 threshold value for this parameter in relation to the potential erosion rate were reported

11 In short, soil response to IE was mainly controlled by the soil aggregates stability
12 against the rainfall erosivity and for the soil tendency to sealing and crusting. These
13 finding has been commonly reported in several studies conducted in similar agricultural
14 areas such as those analyzed here (e.g. Capra *et al.*, 2012; Cerdà 1997, 2002; Ramos *et*
15 *al.*, 2000; Rodrigo Comino *et al.*, 2016b,c). Therefore, the 6 most important soil variables
16 here presented (i.e. E₂, CR₆, CH₃, HY₂, SS₁, and PR₃) would allow identified soil behavior
17 against the IE processes. Moreover, these variables were determined through a quick and
18 easy methodologies (Table S1), which allows a rapid and economic characterization of
19 soil erodibility to IE on large study areas. Nevertheless, due to the high variability in the
20 soil properties that characterize agricultural soils (Ruiz Sinoga & Martinez Murillo, 2009)
21 the guide values of these variables can only applied in similar soils to those evaluated
22 herein. Further investigation is needed to evaluate the 6 keys variables and their
23 interaction on other types of soils. This would allow to define soil erodibility indices that
24 can eventually be applied to improve current empirical models about IE.

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CONCLUSIONS

The soils response to interrill erosion was mainly reflected in 6 variables identified from multivariate statistical procedures: the K factor of the RUSLE model (E_2); an erodibility index obtained from a granulometric balance and the organic matter content in the soil (CR_6); the soil exchangeable sodium percentage (CH_3); the soil seal permeability (HY_2); the soil shear strength (SS_1); and the soil resistance to penetration (PR_3). For each variables, guide values roughly indicating the transition between vulnerable and resistance soils are given. These can be at least successfully applied in some typical Mediterranean soils.

All these variables resulted from field and laboratory techniques, which are economic and easily implemented, allowing then a relatively rapid characterization of large areas, e.g., catchment scale.

The varied nature –physical, chemical and physical-chemical– of the soil properties comprising those variables permits one to assume that they could be also successfully applied even in soils of different natures and characteristics from those studied in this work. However, in this latter case, the guide values proposed –given their empirical character– would probably have to be reformulated. It is advisable to count on techniques permitting the evaluation of soil susceptibility to forming crusts (e.g. many European loess soils) without it being necessarily present at the moment of its determination, since in agricultural soils frequent tilling could be hiding this important problem. A variable for evaluating that soil propensity is proposed (HY_2). Further research is needed to study the interaction among the 6 key variables prosed herein. And after that, to define new empirical models about interrill erosion or to improve current ones.

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SUPPORTING INFORMATION

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Supporting Data

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Data S1. Google Earth file with the location of the 23 studied area.

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Supporting Tables

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REFERENCES

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Akaike H. 1974. A New Look at the Statistical Model Identification. *IEEE Transactions on Automatic Control* **AC-19**: 716-723. DOI: 10.1109/TAC.1974.1100705

Amézketa E. 1999. Soil aggregate stability: A review. *Journal of Sustainable Agriculture* **14**: 83-151.

- 1 Anderberg MR. 1973. Cluster analysis for applications. *Academic Press, New York*: 359
2 pp.
- 3 Andrews SS, Karlen DL, Mitchell JP. 2002. A comparison of soil quality indexing
4 methods for vegetable production systems in Northern California. *Agriculture,*
5 *Ecosystems and Environment* **90**: 25-45. DOI: 10.1016/S0167-8809(01)00174-8.
- 6 Arjmand Sajjadi S, Mahmoodabadi M. 2015. Aggregate breakdown and surface seal
7 development influenced by rain intensity, slope gradient and soil particle size. *Solid*
8 *Earth* **6 (1)**: 311-321. DOI: 10.5194/se-6-311-2015.
- 9 Arnaez J, Lasanta T, Ruiz-Flaño P, Ortigosa L.2007. Factors affecting runoff and erosion
10 under simulated rainfall in Mediterranean vineyards. *Soil and Tillage Research*, **93**
11 **(2)**: 324-334.
- 12 Assouline S, Mualem Y. 2006. Runoff from heterogeneous small bare catchments during
13 soil surface sealing. *Water Resources Research* **42**. DOI: 10.1029/2005WR004592.
- 14 Bagarello V, Ferro V. 2004. Plot-scale measurement of soil erosion at the experimental
15 area of Sparacia (southern Italy). *Hydrological Processes* **18 (1)**: 141-157. DOI:
16 10.1002/hyp.1318.
- 17 Barthès B, Roose E. 2002. Aggregate stability as an indicator of soil susceptibility to
18 runoff and erosion; validation at several levels. *Catena* **47**: 133-149. DOI:
19 10.1016/S0341-8162(01)00180-1.
- 20 Bartlett MS. 1954. A note on the multiplying factors for various chi square
21 approximations. *J. R. Stat. Soc.* **16**: 296-298.
- 22 Ben-Hur M, Agassi M. 1997. Predicting interrill erodibility factor from measured
23 infiltration rate. *Water Resources Research* **33**: 2409-2415.

- 1 Biswas H, Raizada A, Mandal D, Kumar S, Srinivas S, Mishra PK. 2015. Identification
2 of areas vulnerable to soil erosion risk in India using GIS methods. *Solid Earth* **6** (4):
3 1247-1257. DOI: 10.5194/se-6-1247-2015.
- 4 Boardman J, Poesen J. 2006. Soil Erosion in Europe: Major Processes, Causes and
5 Consequences. In *Soil Erosion in Europe*: 477-487. DOI: 10.1002/0470859202.
- 6 Boiffin J, Monnier G. 1985. Infiltration rate as affected by soil surface crusting caused by
7 rainfall. *International symposium on the assessment of soil surface sealing and*
8 *crusting*. In: *Assessment of soil surface sealing and crusting, proceedings of the*
9 *Symposium held in Gent, Belgium*: 210-217.
- 10 Boix-Fayos C, Calvo-Cases A, Imeson AC, Soriano-Soto MD. 2001. Influence of soil
11 properties on the aggregation of some Mediterranean soils and the use of aggregate
12 size and stability as land degradation indicators. *Catena* **44**: 47-67. DOI:
13 10.1016/S0341-8162(00)00176-4.
- 14 Bradford JM, Truman CC, Huang C. 1992. Comparison of three measures of resistance
15 of soil surface seals to raindrop splash. *Soil Technology* **5**: 47-56. DOI:
16 10.1016/0933-3630(92)90006-M.
- 17 Bronick CJ, Lal R. 2005. Soil structure and management: A review. *Geoderma* **124**: 3-
18 22. DOI: 10.1016/j.geoderma.2004.03.005.
- 19 Bryan RB. 2000. Soil erodibility and processes of water erosion on hillslope.
20 *Geomorphology* **32**: 385-415. DOI: 10.1016/S0169-555X(99)00105-1.
- 21 Capra A, Ferro V, Porto P, Scicolone B. 2012. Quantifying interrill and ephemeral gully
22 erosion in a small Sicilian basin. *Zeitschrift fur Geomorphologie* **56**: 9-25. DOI:
23 10.1127/0372-8854/2012/S-00070.

- 1 Casalí J, López JJ, Giráldez JV. 1999. Ephemeral gully erosion in southern Navarra
2 (Spain). *Catena* **36**: 65-84. DOI: 10.1016/S0341-8162(99)00013-2.
- 3 Casalí J, Gastesi R, Álvarez-Mozos J, De Santisteban LM, Lersundi, J D V d, Giménez
4 R, Larrañaga A, Goñi M, Agirre U, Campo MA, López JJ, Donézar M. 2008. Runoff,
5 erosion, and water quality of agricultural watersheds in central Navarre (Spain).
6 *Agricultural Water Management* **95**: 1111-1128. DOI: 10.1016/j.agwat.2008.06.013.
- 7 Cerdà A. 1997. Soil erosion after land abandonment in a semiarid environment of
8 southeastern Spain. *Arid Soil Res. Rehabil.* **11**: 163–176.
9 DOI:10.1080/15324989709381469.
- 10 Cerdà A. 2001. Effects of rock fragment cover on soil infiltration, interrill runoff and
11 erosion. *European Journal of Soil Science* 52 (1): 59-68. DOI:10.1046/j.1365-
12 2389.2001.00354.x.
- 13 Cerdà A, González-Pelayo Ó, Giménez-Morera A, Jordán A, Pereira P, Novara A, Brevik
14 EC, Prosdocimi M, Mahmoodabadi M, Keesstra S, Orenes FG, Ritsema CJ, 2016.
15 Use of barley straw residues to avoid high erosion and runoff rates on persimmon
16 plantations in Eastern Spain under low frequency–high magnitude simulated rainfall
17 events. *Soil Res.* **54**: 154–165.
- 18 Cerdan O, Govers G, Le Bissonnais Y, Van Oost K, Poesen J, Saby N, Gobin A, Vacca
19 A, Quinton J, Auerwald K, Klik A, Kwaad FJPM, Raclot D, Ionita I, Rejman J,
20 Rousseva S, Muxart T, Roxo MJ, Dostal T. 2010. Rates and spatial variations of soil
21 erosion in Europe: a study based on erosion plot data. *Geomorphology* **122**: 167–177.
- 22 Comerma J, Torres S, Lobo D, Fernández N, Delgado R, Madero L. 1992. Aplicación del
23 sistema de evaluación de tierras de la F.A.O. 1985 en la zona de Turén, Venezuela.
24 Cuadernos de Agronomía, año 1 1: 24.

- 1 Dai Q, Liu Z, Shao H, Yang Z. 2015. Karst bare slope soil erosion and soil quality: A
2 simulation case study. *Solid Earth* **6** (3): 985-995. DOI: 10.5194/se-6-985-2015.
- 3 De Ploey J, Múcher HJ. 1981. Consistency index and rainwash mechanisms on belgian
4 loamy soils. *Earth Surface Processes and Landforms* **6**: 319-330.
- 5 Dimoyiannis DG, Tsadilas CD, Valmis S. 1998. Factors affecting aggregate instability of
6 Greek agricultural soils. *Communications in Soil Science and Plant Analysis* **29**:
7 1239-1251.
- 8 FAO. 1980. Metodología provisional para la evaluación de la degradación de los suelos.
9 *F.A.O. 1980, Roma-Italia*: 86.
- 10 Florentino A. 1998. Guía para la evaluación de la degradación del suelo y de la
11 sostenibilidad del uso de la tierra: selección de indicadores físicos. Valores
12 críticos. *En: Manejo Sostenible de los Suelos, Manual de Prácticas. Facultad de*
13 *Agronomía UCV. Maracay- Venezuela*: 68-77.
- 14 Foster, G.R. 1986. Understanding ephemeral gully erosion. Soil conservation, Vol . 2.
15 National Academy of Science Press, Washington, DC,: 90-125.
- 16 García-Ruiz JM, Beguería S, Nadal-Romero E, González-Hidalgo JC, Lana-Renault N,
17 Sansjuan Y. 2015. A meta-analysis of soil erosion rates across the world.
18 *Geomorphology* **239**: 160–173.
- 19 Gessesse B, Bewket W, Bräuning A. 2015. Model-Based Characterization and
20 Monitoring of Runoff and Soil Erosion in Response to Land Use/land Cover Changes
21 in the Modjo Watershed, Ethiopia. *Land Degradation and Development* **26** (7): 711-
22 724. DOI: 10.1002/ldr.2276.
- 23 Giménez R, Casalí J, Grande I, Díez J, Campo MA, Álvarez-Mozos J, Goñi M. 2012.
24 Factors controlling sediment export in a small agricultural watershed in Navarre

1 (Spain). *Agricultural Water Management* **110**: 1-8. DOI:
2 10.1016/j.agwat.2012.03.007.

3 Gumiere SJ, Le Bissonnais Y, Raclot D. 2009. Soil resistance to interrill erosion: Model
4 parameterization and sensitivity. *Catena* **77**: 274-284. DOI:
5 10.1016/j.catena.2009.02.007.

6 Hanson GJ, Cook KR. 2004. Apparatus, test procedures, and analytical methods to
7 measure soil erodibility in situ. *Applied Engineering in Agriculture* **20**: 455-462.

8 Herrick JE, Jones TL. 2002. A dynamic cone penetrometer for measuring soil penetration
9 resistance. *Soil Science Society of America Journal* **66**: 1320-1324.

10 Iserloh T, Ries JB, Arnáez J, Boix-Fayos C, Butzen V, Cerdà A, Echeverría MT,
11 Fernández-Gálvez J, Fister W, Geißler C, Gómez JA, Gómez-Macpherson H, Kuhn
12 NJ, Lázaro R, León FJ, Martínez-Mena M, Martínez-Murillo JF, Marzen M,
13 Mingorance MD, Ortigosa L, Peters P, Regüés D, Ruiz-Sinoga JD, Scholten T,
14 Seeger M, Solé-Benet A, Wengel R, Wirtz S. 2013. European small portable rainfall
15 simulators: A comparison of rainfall characteristics. *Catena* **110**: 100–112.
16 DOI:10.1016/j.catena.2013.05.013.

17 IUSS Working Group WRB. 2014. World Reference Base for Soil Resources. World Soil
18 Resources Report, FAO, Rome.

19 Keesstra SD, Maroulis J, Argaman E, Voogt A, Wittenberg L. 2014. Effects of controlled
20 fire on hydrology and erosion under simulated rainfall. *Cuadernos de Investigación*
21 *Geográfica* **40**: 269-293. DOI: 10.18172/cig.2532.

22 Kabacoff R. 2015. R in Action: Data Analysis and Graphics with R. *Manning*
23 *Publications Co. Greenwich, CT, USA*: 475 pp.

- 1 Kaiser HF. 1958. The varimax criterion for analytic rotation in factor analysis.
2 *Psychometrika* **23**: 187-200. DOI: 10.1007/BF02289233.
- 3 Kamphorst A. 1987. A small rainfall simulator for the determination of soil erodibility.
4 *Netherlands Journal of Agricultural Science* **35**: 407–415.
- 5 Kemper WD, Koch EJ. 1966. Aggregate stability of soils from western USA and Canada.
6 *US Government Printing Office, Washington, DC. USDA Technical Bulletin* **1355**:
7 52 pp.
- 8 Laflen JM. 2003. Erosion Process Modeling. In: *Encyclopedia of Water Science*: 218-
9 221. DOI: 10.1081/E-EWS-120010224.
- 10 Lassu T, Seeger M, Peters P, Keesstra SD. 2015. The Wageningen Rainfall Simulator:
11 Set-up and Calibration of an Indoor Nozzle-Type Rainfall Simulator for Soil Erosion
12 Studies. *Land Degrad. Dev.* **26**: 604–612. DOI:10.1002/ldr.2360.
- 13 Laufer D, Loibl B, Märlander B, Koch HJ. 2016. Soil erosion and surface runoff under
14 strip tillage for sugar beet (*Beta vulgaris* L.) in Central Europe. *Soil and Tillage*
15 *Research* **162**: 1-7. DOI: 10.1016/j.still.2016.04.007.
- 16 Le Bissonnais Y. 1996. Aggregate stability and assessment of soil crustability and
17 erodibility: I. Theory and methodology. *European Journal of Soil Science* **47**: 425-
18 437.
- 19 Le Bissonnais Y, Arrouays D. 1997. Aggregate stability and assessment of soil
20 crustability and erodibility: II. Application to humic loamy soils with various organic
21 carbon contents. *European Journal of Soil Science* **48**: 39-48.
- 22 Le Bissonnais Y, Cerdan O, Lecomte V, Benkhadra H, Souchère V, Martin P. 2005.
23 Variability of soil surface characteristics influencing runoff and interrill erosion.
24 *Catena* **62**: 111-124. DOI: 10.1016/j.catena.2005.05.001.

- 1 Leguédou S, Le Bissonnais Y. 2004. Size fractions resulting from an aggregate stability
2 test, interrill detachment and transport. *Earth Surface Processes and Landforms* **29**
3 **(9)**: 1117-1129. DOI: 10.1002/esp.1106.
- 4 Léonard J, Richard G. 2004. Estimation of runoff critical shear stress for soil erosion from
5 soil shear strength. *Catena* **57**: 233-249. DOI: 10.1016/j.catena.2003.11.007.
- 6 Lin F. 2008. Solving multicollinearity in the process of fitting regression model using the
7 nested estimate procedure. *Quality and Quantity* **42**: 417-426. DOI: 10.1007/s11135-
8 006-9055-1.
- 9 Lo, shuh-shiaw. 1992. Glossary of hydrology. Water Resources Publications, Colorado
10 80161, USDA. pp1793
- 11 Luk S, Hamilton H. 1986. Experimental effects of antecedent moisture and soil strength
12 on rainwash erosion of two luvisols, Ontario. *Geoderma* **37**: 29-43. DOI:
13 10.1016/0016-7061(86)90041-8. DOI: 10.1016/0016-7061(86)90041-8.
- 14 Ma RM, Li ZX, Ca CF, Wang JG. 2014. The dynamic response of splash erosion to
15 aggregate mechanical breakdown through rainfall simulation events in Ultisols
16 (subtropical China). *Catena* **121**: 279-287. DOI: 10.1016/j.catena.2014.05.028.
- 17 Maetens W, Vanmaercke M, Poesen J, Jankauskas B, Jankauskiene G, Ionita I. 2012.
18 Effects of land use on annual runoff and soil loss in Europe and the Mediterranean:
19 a meta-analysis of plot data. *Prog. Phys. Geogr.* 1–55.
20 <http://dx.doi.org/10.1177/0309133312451303>.
- 21 Marzen M, Iserloh T, Casper MC, Ries JB. 2015. Quantification of particle detachment
22 by rain splash and wind-driven rain splash. *Catena* **127**: 135-141.

- 1 Mamedov AI, Shainberg I, Levy GJ. 2002. Wetting rate and sodicity effects on interrill
2 erosion from semi-arid Israeli soils. *Soil and Tillage Research* 68: 121-132. DOI:
3 10.1016/S0167-1987(02)00115-0.
- 4 Martínez-Zavala L, Jordán A. 2008. Effect of rock fragment cover on interrill soil erosion
5 from bare soils in Western Andalusia, Spain. *Soil Use and Management* **24 (1)**: 108-
6 117.
- 7 Meyer LD, Harmon WC. 1984. Susceptibility of agricultural soils to interrill erosion. *Soil*
8 *Science Society of America Journal* **48**: 1152-1157.
- 9 Moriasi DN, Arnold JG, Van Liew MW, Bingner RL, Harmel RD, Veith TL. 2007. Model
10 evaluation guidelines for systematic quantification of accuracy in watershed
11 simulations. *Transactions of the ASABE* **50**: 885-900.
- 12 Nash JE, Sutcliffe JV. 1970. River flow forecasting through conceptual models part I - A
13 discussion of principles. *Journal of Hydrology* **10**: 282-290. DOI: 10.1016/0022-
14 1694(70)90255-6.
- 15 Neave M, Rayburg S. 2007. A field investigation into the effects of progressive rainfall-
16 induced soil seal and crust development on runoff and erosion rates: The impact of
17 surface cover. *Geomorphology* **87 (4)**: 378-390. DOI:
18 10.1016/j.geomorph.2006.10.007.
- 19 Ola A, Dodd IC, Quinton JN. 2015. Can we manipulate root system architecture to control
20 soil erosion? *SOIL* **1**: 603-612. DOI: 10.5194/soil-1-603-2015.
- 21 Panagos P, Meusburger K, Ballabio C, Borrelli P, Alewell C. 2014. Soil erodibility in
22 Europe: A high-resolution dataset based on LUCAS. *Science of the Total*
23 *Environment* **479-480**: 189-200. DOI: 10.1016/j.scitotenv.2014.02.010

- 1 Pla I. 1982. Sealing index to predict problems of soil and water conservation in tropical
2 rainfed agricultural lands. *ASAESSASSSA, Annual Meetings. Anaheim, California.*
3 *USA. 1982.*
- 4 Poesen JW, Torri D, Bunte K. 1994. Effects of rock fragments on soil erosion by water
5 at different spatial scales: a review. *Catena* **23**: 141-166. DOI: 10.1016/0341-
6 8162(94)900.
- 7 Poesen J, Lavee H. 1994. Rock fragments in top soils: significance and processes. *Catena*
8 **23**: 1–28. doi:10.1016/0341-8162(94)90050-7.
- 9 Prosdocimi M, Burguet M, Di Prima S, Sofia G, Terol E, Rodrigo Comino J, Cerdà A,
10 Tarolli P. 2016a. Rainfall simulation and Structure-from-Motion photogrammetry
11 for the analysis of soil water erosion in Mediterranean vineyards. *Sci. Total Environ.*
12 **574**: 204–215. DOI:10.1016/j.scitotenv.2016.09.03658-2.
- 13 Prosdocimi M, Cerdà A, Tarolli P. 2016b. Soil water erosion on Mediterranean vineyards:
14 A review. *Catena* **141**: 1–21. DOI:10.1016/j.catena.2016.02.010.
- 15 Pulido-Moncada MA, Lobo-Luján D, Lozano-Prez Z. 2009. Association between soil
16 structure stability indicators and organic matter in Venezuelan agricultural soils.
17 *Agrociencia* **43**: 221-230.
- 18 R Core Team. 2015. R: A language and environment for statistical computing. R
19 Foundation for Statistical Computing, Vienna, Austria. URL: [http://www.R-](http://www.R-project.org/)
20 [project.org/](http://www.R-project.org/).
- 21 Ramos MC, Nacci S, Pla I. 2003. Effect of raindrop impact and its relationship with
22 aggregate stability to different disaggregation forces. *Catena* **53**: 365-376. DOI:
23 10.1016/S0341-8162(03)00086-9.

- 1 Ramos MC, Martínez-Casasnovas JA. 2007. Soil loss and soil water content affected by
2 land levelling in Penedès vineyards, NE Spain. *Catena* **71**: 210–217. DOI:
3 <http://dx.doi.org/10.1016/j.catena.2007.03.001>.
- 4 Rodrigo Comino J, Iserloh T, Morvan X, Malam Issa O, Naisse C, Keesstra SD, Cerdà
5 A, Prosdocimi M, Arnáez J, Lasanta T, Ramos MC, Marqués MJ, Ruiz Colmenero
6 M, Bienes R, Ruiz Sinoga JD, Seeger M, Ries JB. 2016b. Soil Erosion Processes in
7 European Vineyards: A Qualitative Comparison of Rainfall Simulation
8 Measurements in Germany, Spain and France. *Hydrology* **3**, **6**.
9 DOI:10.3390/hydrology3010006.
- 10 Rodrigo Comino J, Ruiz Sinoga JD, Senciales González JM, Guerra-Merchán A, Seeger
11 M, Ries JB, 2016c. High variability of soil erosion and hydrological processes in
12 Mediterranean hillslope vineyards (Montes de Málaga, Spain). *Catena* **145**: 274–284.
13 DOI:10.1016/j.catena.2016.06.012.
- 14 Romero CC, Stroosnijder L, Baigorria GA. 2007. Interrill and rill erodibility in the
15 northern Andean Highlands. *Catena***70**: 105-113. DOI:
16 [10.1016/j.catena.2006.07.005](http://dx.doi.org/10.1016/j.catena.2006.07.005).
- 17 Römken MJM. 1985. The soil erodibility factor: a perspective. *Soil erosion and*
18 *conservation*: 445-461.
- 19 Ruiz Sinoga JD, Martínez Murillo JF, 2009. Hydrological response of abandoned
20 agricultural soils along a climatological gradient on metamorphic parent material in
21 southern Spain. *Earth Surf. Process. Landf.* **34**: 2047–2056. DOI:10.1002/esp.1890.
- 22 Shainberg I, Letey J. 1984. Response of soils to sodic and saline conditions. *Hilgardia*
23 **52**: 1-57.

- 1 Sheridan GJ, So HB, Loch RJ, Walker CM. 2000. Estimation of erosion model erodibility
2 parameters from media properties. *Australian Journal of Soil Research* **38**: 265-284.
3 DOI: 10.1071/SR99041.
- 4 Shrestha S, Kazama F. 2007. Assessment of surface water quality using multivariate
5 statistical techniques: A case study of the Fuji river basin, Japan. *Environmental*
6 *Modelling and Software* **22**: 464-475. DOI: 10.1016/j.envsoft.2006.02.001.
- 7 Singer MJ, Janitzky P, Blackard J. 1982. The influence of exchangeable sodium
8 percentage on soil erodibility. *Soil Science Society of America Journal* **46**: 117-121.
- 9 Truman CC, Bradford JM. 1990. Effect of antecedent soil moisture on splash detachment
10 under simulated rainfall. *Soil Science* **150**: 787-798.
- 11 Vandekerckhove J, Matzke D, Wagenmakers EJ. 2015. Model comparison and the
12 principle of parsimony. In Busemeyer JR, Townsend JT, Wang ZJ, Eidels A. (Eds.)
13 *Oxford Handbook of Computational and Mathematical Psychology*: 300-317.
14 Oxford, UK: Oxford University Press.
- 15 Vega M, Pardo R, Barrado E, Debán L. 1998. Assessment of seasonal and polluting
16 effects on the quality of river water by exploratory data analysis. *Water research* **32**:
17 3581-3592. DOI: 10.1016/S0043-1354(98)00138-9.
- 18 Verhaegen T. 1984. The influence of soil properties on the erodibility of Belgian loamy
19 soils: a study based on rainfall simulation experiments. *Earth Surface Processes &*
20 *Landforms* **9**: 499-507.
- 21 Walpole RE, Myers RH, Myers SL, Ye K. 2012. Probability and Statistics for Engineers
22 and Scientists (9th Edition). *Pearson Prentice Hall*: 816 pp.

- 1 Wang G, Fang Q, Wu B, Yang H, Xu Z. 2015. Relationship between soil erodibility and
2 modeled infiltration rate in different soils. *Journal of Hydrology* **528**: 408-418. DOI:
3 10.1016/j.jhydrol.2015.06.044.
- 4 Wang Y, Zhang JH, Zhang ZH. 2015. Influences of intensive tillage on water-stable
5 aggregate distribution on a steep hillslope. *Soil and Tillage Research* **151**: 82-92.
6 DOI: 10.1016/j.still.2015.03.003.
- 7 Ward JH. 1963. Hierarchical grouping to optimize an objective function. *Journal of the*
8 *American Statistical Association* **58**: 236-244. DOI:
9 10.1080/01621459.1963.10500845.
- 10 Westra S, Brown C, Lall U, Koch I, Sharma A. 2010. Interpreting variability in global
11 SST data using independent component analysis and principal component analysis.
12 *International Journal of Climatology* **30**: 333-346. DOI: 10.1002/joc.1888.
- 13 Wilcox BP, Wood MK. 1989. Factors influencing interrill erosion from semiarid slopes
14 in New Mexico. *Journal of Range Management* **42**: 66-70.
- 15 Willet P. 1987. Similarity and Clustering in Chemical Information Systems. *Research*
16 *Studies Press, Wiley, New York*: 253 pp.
- 17 Wischmeier WH, Smith DD. 1978. Predicting rainfall erosion losses - A guide to
18 conservation planning. *USDA-Agr. Handbook* **537**: 65 pp.

1 Tables

2 Table I. Main physical-chemical properties of the soils (epipedon) in the study. EC =
 3 electrical conductivity, ESP = exchange sodium percentage, CEC = cation exchange
 4 capacity.

Soil name	Soil type ^a (FAO-WRB)	Sand (%)	Silt (%)	Clay (%)	Organic matter (%)	Stoniness (%)	Bulk density (g cm ⁻³)	EC (dS m ⁻¹)	ESP (%)	pH	CEC (cmol (+) kg ⁻¹)	CaCO ₃ (%)
PIT 1	Pellic cambisol	26.2	48.5	25.3	1.1	0.6	1.6	7.9	66.5	8.9	16.0	41.2
PIT 2	Pellic cambisol	10.7	62.4	26.9	1.7	7.0	1.5	21.9	72.0	8.1	12.0	38.9
PIT 3	Pellic cambisol	30.7	44.8	24.4	1.0	2.2	1.4	0.7	8.4	8.5	10.0	40.9
AOI 1	Rendzic cambisol	5.5	61.9	32.6	3.3	12.1	1.2	0.6	1.8	8.1	16.8	39.4
AOI 2	Rendzic cambisol	10.4	63.1	26.5	2.3	25.5	1.2	0.5	3.0	8.1	13.0	46.8
AOI 3	Rendzic cambisol	13.8	48.8	37.4	2.3	19.5	1.2	0.3	1.6	8.3	16.3	25.2
AOI 4	Rendzic cambisol	17.3	49.2	33.5	1.9	26.9	1.3	0.6	2.1	8.3	14.3	25.4
AOI 5	Rendzic cambisol	20.3	44.6	35.1	2.2	17.6	1.4	0.8	3.4	8.5	14.8	24.9
AOI 6	Rendzic cambisol	66.4	21.8	11.8	1.1	0.6	1.5	0.3	2.0	7.8	6.4	0.6
LUM 1	Pellic kastanozen	13.1	55.4	31.5	2.3	12.6	1.2	0.3	3.8	7.3	14.6	40.1
LUM 2	Pellic kastanozen	9.4	52.8	37.8	1.8	2.7	1.4	0.3	3.2	8.3	15.5	39.3
LUM 3	Pellic kastanozen	8.2	65.7	26.1	2.3	1.9	1.4	0.4	4.1	8.3	13.0	46.2
ABA 1	Rendzic leptosol	17.3	55.1	27.5	4.3	30.2	1.1	0.6	1.5	7.2	17.0	1.0
ABA 2	Rendzic leptosol	24.2	47.3	28.6	3.7	5.1	1.2	0.7	0.9	7.0	15.5	1.1
LEO 1	Haplic leptosol	28.8	45.1	26.1	1.4	0.8	1.5	0.4	1.2	7.9	11.2	1.3
LEO 2	Haplic leptosol	52.0	32.9	15.2	2.3	44.5	1.1	0.8	1.3	7.6	10.8	3.5
RAD 1	Eutric cambisol	6.9	55.1	38.0	1.1	1.1	1.1	3.2	7.1	7.7	14.3	7.8
RAD 2	Eutric cambisol	12.5	60.8	26.8	1.2	25.9	1.2	3.6	12.6	7.8	11.1	12.5
RAD 3	Eutric cambisol	6.6	42.0	51.4	1.6	6.9	1.0	1.7	12.0	8.2	19.1	8.5
RAD 4	Eutric cambisol	8.4	52.6	38.9	2.0	17.9	1.1	1.0	8.2	8.3	16.1	10.7
RAD 5	Eutric cambisol	36.5	42.9	20.6	1.1	41.7	1.1	4.3	27.4	7.8	9.0	1.3
RAD 6	Eutric cambisol	28.2	24.7	47.1	1.6	20.9	1.1	1.2	4.6	7.7	17.8	0.8
RAD 7	Eutric cambisol	21.3	50.0	28.7	1.5	64.7	1.2	9.8	31.1	7.8	12.3	32.5

5 ^a: soil type classification according IUSS (2014).

6 Table II. Inset of the 2 clusters obtained from the cluster analysis (CA).

CLUSTER 1								
Variables ^a	t-test ^b	Significance level	p value	Mean in cluster	% Variation	Overall mean	SD ^c in cluster	Overall SD ^c
CH ₅	-3.97	****	0.0001	10.79	-20.15	13.52	1.84	3.08
CR ₅	3.82	****	0.0001	1.20	24.29	0.97	0.16	0.28
SI ₂	3.79	****	0.0002	2.23	23.83	1.80	0.24	0.51
SI ₃	3.59	****	0.0003	2.39	23.59	1.94	0.31	0.57
E ₂	3.27	****	0.001	0.04	28.71	0.03	0.01	0.01
CR ₄	2.98	****	0.003	4.10	31.32	3.12	1.41	1.48
CR ₃	2.96	****	0.003	3.22	35.22	2.38	1.37	1.28
PR ₃	2.87	****	0.004	900.59	35.77	663.30	413.58	370.85
GT ₁	2.70	***	0.007	0.001	35.27	0.0004	0.0003	0.0002
PR ₂	2.56	**	0.013	787.18	31.87	596.94	394.69	334.30
CR ₁	2.47	**	0.014	0.59	23.80	0.47	0.22	0.21
HY ₂	-2.39	**	0.017	1.74	-20.77	2.19	0.47	0.85
CR ₆	2.37	**	0.018	1.01	16.21	0.87	0.27	0.27
GF ₁	2.31	*	0.021	0.00	56.57	0.00	0.00	0.00
CH ₁	-2.24	*	0.025	1.53	-21.68	1.96	0.51	0.85
PH ₁	2.13	*	0.033	1.32	5.91	1.25	0.16	0.16
GT ₄	2.05	*	0.041	0.04	62.69	0.02	0.04	0.03
CH ₃	2.03	*	0.042	20.87	71.56	12.17	24.85	19.24
HY ₂	-2.00	*	0.046	1.84	-20.03	2.30	0.85	1.04
CLUSTER 2								
CH ₅	3.97	****	0.0001	16.02	18.47	13.52	1.43	3.08
CR ₅	-3.82	****	0.0001	0.75	-22.26	0.97	0.16	0.28
SI ₂	-3.79	****	0.0002	1.41	-21.84	1.80	0.35	0.51
SI ₃	-3.59	****	0.0003	1.52	-21.62	1.94	0.42	0.57
E ₂	-3.27	****	0.001	0.02	-24.74	0.03	0.01	0.01
CR ₄	-2.98	****	0.003	2.23	-28.71	3.12	0.81	1.48
CR ₃	-2.96	****	0.003	1.61	-32.29	2.38	0.39	1.28
PR ₃	-2.87	****	0.004	445.79	-32.79	663.30	88.80	370.85
GT ₁	-2.70	***	0.007	0.0002	-32.33	0.0004	0.0001	0.0002
PR ₂	-2.56	**	0.013	422.54	-29.22	596.94	88.37	334.30
CR ₁	-2.47	**	0.014	0.37	-21.81	0.47	0.11	0.21
HY ₂	2.39	**	0.017	2.61	19.03	2.19	0.91	0.85
CR ₆	-2.37	**	0.018	0.74	-14.86	0.87	0.19	0.27
GF ₁	-2.31	*	0.021	0.00	-51.85	0.00	0.00	0.00
CH ₁	2.24	*	0.025	2.35	19.87	1.96	0.91	0.85
PH ₁	-2.13	*	0.033	1.18	-5.42	1.25	0.11	0.16
GT ₄	-2.05	*	0.041	0.01	-57.47	0.02	0.00	0.03
CH ₃	-2.03	*	0.042	4.19	-65.59	12.17	3.19	19.24
HY ₂	2.00	*	0.046	2.73	18.36	2.30	1.02	1.04

7 * p < 0.05; ** p < 0.02; *** p < 0.01; **** p < 0.005.

8 a: the names and the measure units of the soil parameters are displayed in Table S1.

9 b: value of Student's t-test.

10 c: standard deviation.

- 1 Table III. Principal components (PCs) with an eigenvalue value of above 1 and correlation
- 2 coefficients of the experimental variables over the first two PCs for the 23 soils.

<i>Number of PCs</i>	<i>Eigenvalue</i>	<i>Percentage</i>	<i>Cumulated Percentage</i>
1	10.39	21.65	21.65
2	8.05	16.77	38.42
3	6.12	12.75	51.17
4	4.99	10.39	61.56
5	3.63	7.55	69.11
6	2.95	6.14	75.24
7	2.57	5.36	80.60
8	1.79	3.72	84.32
9	1.54	3.20	87.53
10	1.26	2.63	90.15
11	1.07	2.22	92.38

Active variables – Factors correlations ^a

<i>Variable ^b</i>	<i>Axis 1</i>	<i>Axis 2</i>
CH ₂	0.55	-0.08
CH ₃	0.58	-0.06
CH ₅	-0.38	-0.73
CH ₆	-0.34	0.75
CR ₃	0.13	0.97
CR ₅	0.77	0.35
CR ₆	0.81	-0.22
E ₂	0.86	0.15
GF ₁	0.08	0.95
GF ₂	-0.15	0.92
GF ₃	-0.13	0.95
GF ₄	-0.14	0.97
GF ₅	-0.20	0.82
GT ₁	0.01	0.63
HY ₂	-0.62	0.03
PH ₁	0.62	0.30
PR ₂	0.51	0.14
PR ₃	0.51	0.14
SI ₁	0.69	-0.39
SI ₂	0.63	0.69
SI ₃	0.57	0.71
SS ₁	-0.53	-0.10

Supplementary variables (SL₁) – Factors correlations

<i>Variable</i>	<i>Axis 1</i>	<i>Axis 2</i>
SL ₁	0.35	-0.19

- 3 a: bold factors present a correlation loading greater than 0.50.
- 4 b: the names and the measure units of the soil parameters are displayed in Table S1.

- 1 Table IV. The best linear regression models obtained for one and two independent
 2 variables in terms of the criteria established for regression diagnosing.

Indicators of the regression diagnosis		Model 1	Model 2
Number of variables		1	2
Variables ^a		CH ₃	CH ₃ , CR ₆
NSE ^b		0.915	0.996
MSE ^c		0.04	0.01
RMSE ^d		0.21	0.10
AIC ^e		3.18	-9.30
Regression coefficients	Intercept	0.903	0.813
	1 [*]	1.166	0.151
	2 ^{**}	-	0.575
p (t-test) ^f	Intercept	4.16 e ⁻⁰⁵	2.010 e ⁻⁰⁶
	1 [*]	1.45 e ⁻⁰⁵	4.96 e ⁻⁰⁹
	2 ^{**}	-	2.08 e ⁻⁰⁶
VIF ^g	1 [*]	-	1.05
	2 ^{**}	-	1.05

- 3 *: values for the first variable introduced by the model; **: Values for the second variable introduced by the model.
 4 a: the names and the measure units of the soil parameters are displayed in Table S1.
 5 b: Nash-Sutcliffe efficiency coefficient.
 6 c: mean square error.
 7 d: root mean square error.
 8 e: Akaike information criterion.
 9 f: significance level of Student's t-test.
 10 g: variance inflation factor.

- 1 Table V. Range of the values in the 2 clusters for the 6 variables identified as conditioners
 2 of the interrill erosion after carrying out the 3 multivariate statistics analyses (see caption
 3 in Table S1).

Cluster 1	E₂ (t ha h ha⁻¹ MJ⁻¹mm⁻¹)	CR₆ (dimensionless)	CH₃ (%)	PR₃ (kPa)	SS₁ (kPa)	HY₂ (mm h⁻¹)
Min ^a	0.023	0.455	1.16	194.500	8.728	0.967
Mean ^b	0.040	1.011	20.87	916.153	14.118	1.737
Max ^c	0.053	1.285	71.95	1797.500	20.947	2.506
SD ^d	0.009	0.284	26.06	427.681	2.993	0.495
Cluster 2						
Min ^a	0.015	0.273	0.88	316.000	11.346	1.854
Mean ^b	0.023	0.741	4.19	462.458	16.001	2.610
Max ^c	0.033	0.982	11.98	834.000	27.493	5.399
SD ^d	0.006	0.195	3.33	137.041	4.731	0.952

- 4 a: minimum value of the variable in the cluster.
 5 b: mean value of the variable in the cluster.
 6 c: maximum value of the variable in the cluster.
 7 d: standard deviation of the variable in the cluster.

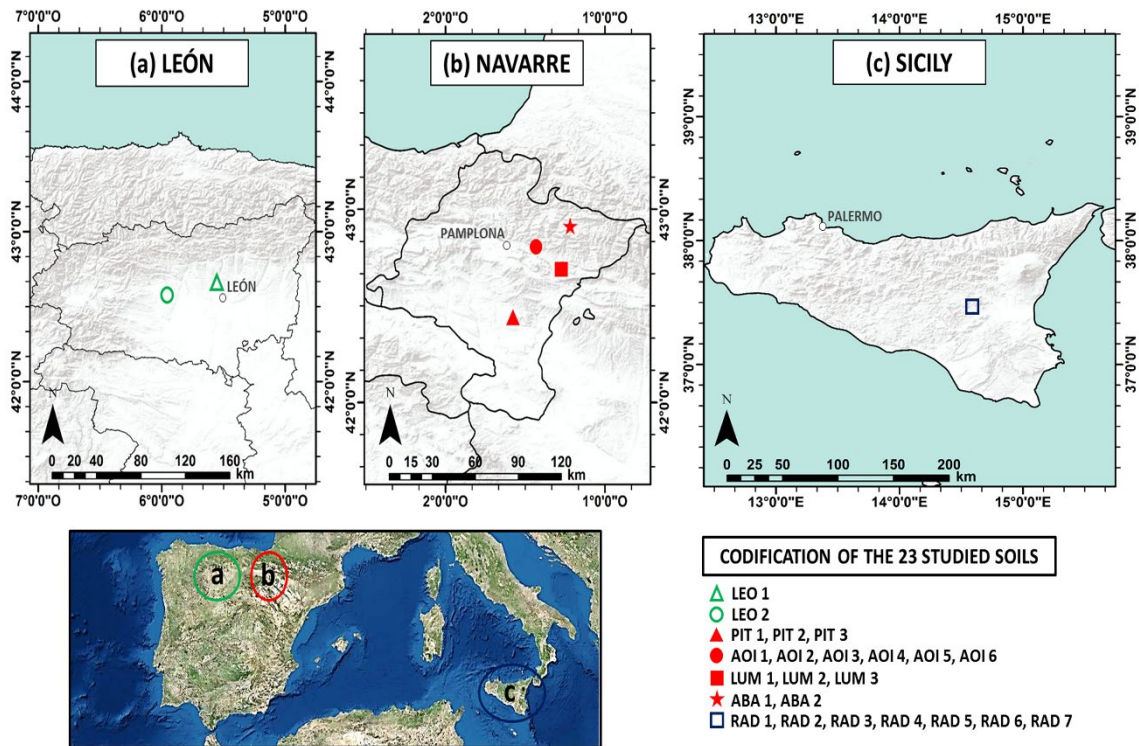
1 **List of figures and titles**

2 Figure 1. Location map of the studied areas: (a) 2 soils in León (northwest of Spain)
3 (green symbols), (b) 14 soils in Navarre (north of Spain) (red symbols), and (c) 7 soils in
4 Sicily (south of Italy) (blue symbol).

5 Figure 2. Rainfall simulator setup: (a) typical experimental site in the field (red arrows),
6 and (b) rainfall simulator.

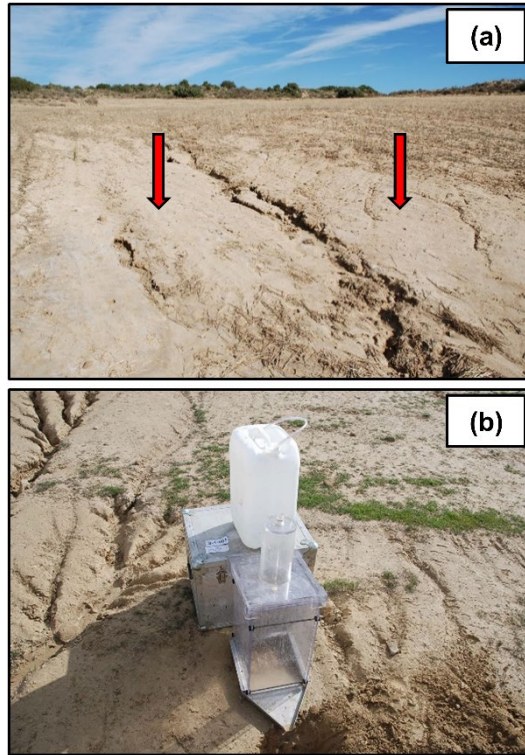
7 Figure 3. Distribution of the mean values and standard deviations for the interrill erosion
8 rate (SL_1) and the 6 soil variables statistically identified in Cluster 1 (red circles and black
9 error bars) and in Cluster 2 (blue squares and green error bars). The gray dashed line
10 demarcates the transition area between the 2 clusters. Variables E_2 (K factor of RUSLE),
11 CH_3 (exchangeable sodium percentage), PR_3 (penetration resistance in the first 6
12 centimeters of the soil depth) and CR_6 (modified crusting index) are positively related to
13 the erosion rate, whereas variables SS_1 (shear strength) and HY_2 (hydraulic conductivity
14 of seal) are negatively related to the aforementioned (e.g. a value higher than 8% for CH_3
15 would imply a higher vulnerability degree of the soil against erosion). The horizontal axis
16 shows the name of the variables. The left-hand vertical axis shows, first, PR_3 (kPa) and
17 SS_1 (kPa); and, second, CH_3 (%). The right-hand vertical axes show first SL_1 ($t\ ha^{-1}\ h^{-1}$),
18 HY_2 ($mm\ h^{-1}$) and CR_6 (dimensionless $\times 10^{-1}$); and, second, E_2 ($t\ ha\ h\ ha^{-1}\ MJ^{-1}\ mm^{-1}$).

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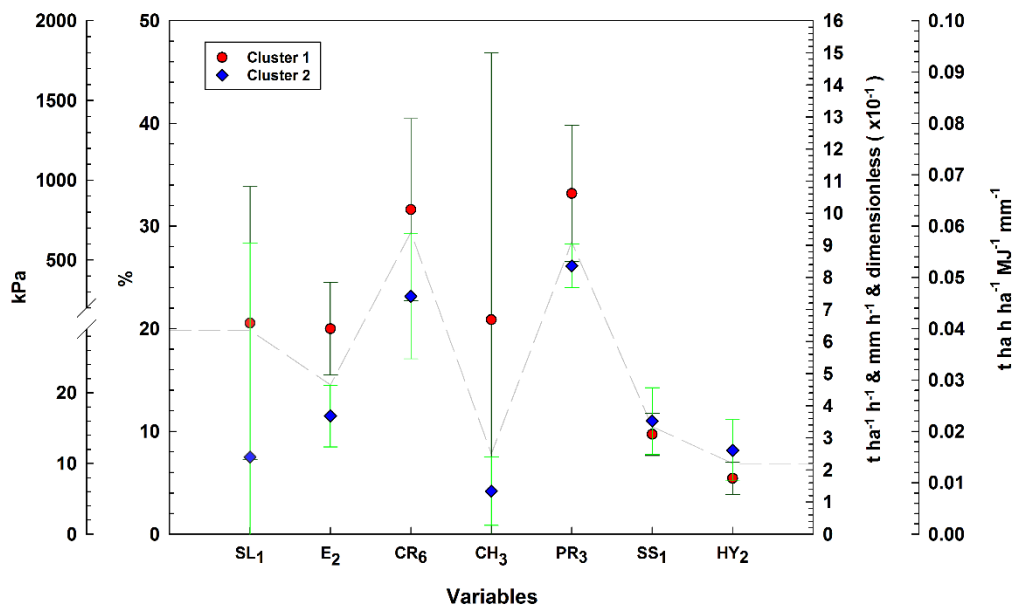
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13 HY₂ (mm h⁻¹) and CR₆ (dimensionless x 10⁻¹); and, second, E₂ (t ha h ha⁻¹ MJ⁻¹ mm⁻¹).